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1 **Wastewater used for urban agriculture in West Africa as a reservoir for**
2 **antibacterial resistance dissemination**

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12
13 **Summary**

14 State of art metagenomics were used to investigate the microbial population, antibiotic resistance genes and
15 plasmids of medical interest in wastewater used for urban agriculture in Ouagadougou (Burkina Faso).

16 Wastewater samples were collected from three canals near agricultural fields in three neighbourhoods.

17 Assessment of microbial population diversity revealed different microbial patterns among the different samples.

18 Sequencing reads from the wastewaters revealed different functional specializations of microbial communities,
19 with the predominance of carbohydrates and proteins metabolism functions. Eleven pathogen-specific and 56

20 orthologous virulence factor genes were detected in the wastewater samples. These virulence factors are usually

21 found in human pathogens that cause gastroenteritis and/or diarrhoea. A wide range of antibiotic resistance

22 genes was identified; 81 are transmissible by mobile genetic elements. These included seven different extended

23 spectrum β -lactamase genes encoding synthesis of four enzyme families, including two metallo- β -lactamases

24 (*bla*_{AIM-1} and *bla*_{GES-21}). Ten different incompatibility groups of *Enterobacteriaceae* plasmid replicons (ColE,

25 FIB, FIC, FII, P, Q, R, U, Y, and A/C), and 30 plasmid replicon types from Gram-positive bacteria. All are

26 implicated in the wide distribution of antibiotic resistance genes. We conclude that wastewater used for urban

27 agriculture in the city represents a high risk for spreading bacteria and antimicrobial resistance among humans

28 and animals.

29 **Keywords:** Antimicrobial resistance, metagenomics, urban agriculture, low and middle-income countries

30

31 **Introduction**

32 There is an increasing number of bacteria that are multi-resistant against common antibiotics and thus
33 cannot be treated by current therapies (Leopold et al. 2014). Antibiotic resistance has led to the need for more
34 expensive drugs, which many cannot afford, resulting in increased morbidity and mortality (Laxminarayan et al.
35 2016). Selection of drug-resistant bacteria and proliferation occurs at sub-inhibitory concentrations of antibiotics
36 (Drlica, 2003; Andersson and Hughes, 2010). It has been reported that approximately 50 - 90 % of antibiotics
37 administered to humans and animals are excreted via urine and faeces, as a mixture of parent drug and
38 metabolite forms, and thus significant levels of active drug end up in the environment, where they may persist in
39 soil and aquatic ecosystems (Kümmerer, 2009). To combat the presence of antibiotics in ecosystems, bacteria
40 have evolved a plethora of different resistance genes of which many are mobile and can easily spread between
41 species including human pathogens (Groh et al. 2007; Wellington et al. 2013; Zhang et al. 2016). Consequently,
42 environmental bacteria serve as a reservoir of resistance genes that can be transmitted to pathogenic species
43 (Allen et al. 2009). That is of concern in low and middle-income countries (LMICs) where populations are often
44 directly exposed to untreated wastewater due to a lack of water and sanitation services including in urban areas
45 (UN, 2017; Bougnom and Piddock, 2017).

46 The urban population in sub-Saharan Africa is projected to grow by 3.5 % per annum from now until
47 2030, and the number of urban residents is expected to rise from 400 million in 2010 to 1.26 billion in 2050
48 (UN, 2012). This rapid increase in the urban population raises significant challenges, related to urban poverty,
49 public health, housing, food security and environmental pollution. Urban and peri-urban agriculture is defined
50 as the production of crops and livestock within and around cities (UNDP, 1996). It has been developed by urban
51 dwellers to meet their food demand, as well as a source of employment and income. The importance of urban
52 agriculture in terms of contributing to food security and poverty alleviation has been recognised and promoted
53 (Martellozzo et al. 2014). Two hundred million urban dwellers are reported to be engaged in urban agriculture
54 worldwide, producing in some cases up to 90 % of cities' demand for perishable vegetables (UN, 2013). It is
55 reported that more than 80% of the domestic and industrial wastewater generated in LMICs is discharged
56 untreated into the environment, and because of its low cost, availability, and nutrient content, urban agriculture
57 hugely relies on wastewater for irrigation, (Mateo-Sagasta et al. 2013). As water is a vehicle for microbial
58 dissemination, wastewater for agricultural irrigation represents a very serious health risk, not least as it increases
59 exposure to faecal pathogens. Ingestion of faecally-contaminated water and/or food by opportunistic and/or
60 pathogenic microorganisms is one of the major reasons for the high number of water-related diseases in LMICs.

61 Wastewater use for urban agriculture has been shown to expose humans and animals to enteric diseases caused
62 by pathogenic bacteria, protozoa, and helminths (Dickin et al. 2016). To reduce the spread of antimicrobial
63 resistant bacteria to humans, the role of wastewater, sanitation and hygiene must be understood and addressed
64 (Graham et al. 2011; Wuijts et al. 2017).

65 Genomics allows us to understand the factors that drive ARGs transfer between bacteria and the
66 mechanisms conferring resistance (Amos et al. 2014). The use of a metagenomic approach provides an
67 opportunity to provide essential information so that the risks of untreated wastewater and impact upon human
68 and animal health can be defined. In this study, we used metagenomics to assess the presence of virulence factor
69 genes (VFs), antibiotic resistance genes (ARGs) and plasmids of medical interest in wastewater used for urban
70 agriculture in Ouagadougou (Burkina Faso). The aim was to measure the potential of urban agriculture to
71 disseminate antibiotic resistance in the city.

72

73 **Material and Methods**

74 *Wastewater sampling*

75 Experiments were conducted with samples collected from Ouagadougou (Burkina Faso, West Africa)
76 in October 2015, at the end of the rainy season, with three precipitation events during the month. Ouagadougou
77 is the capital city of Burkina Faso, it occupies a surface area of 219.3 km², with a population of about 2.7
78 million inhabitants. The city is in the Sudano-Sahelian area, with a rainfall of about 800 mm per year, and
79 temperatures ranging from 16 °C to 43 °C during the rainy and dry seasons, respectively. Wastewater samples
80 used for urban agriculture were collected in three neighborhoods from three canals near agricultural fields. The
81 sampling sites were selected based on their different characteristics. These canals are open-air water drainage
82 and collection points of different transects (Figure 1). They were constructed to protect the city from floods, and
83 receive hand-draining sludge, solid waste, and wastewater. They do not have regular cleaning and are used to
84 irrigate agricultural fields which are found along its way. Passpanga canal is 5 km long, it receives effluent from
85 the university teaching hospital, the city and biggest market of the town; these are the main sources of pollution.
86 Its outlet is dam N° 3. Zogona canal, which is 4.4 km long; it receives effluent from 1200 houses and the
87 University of Ouagadougou as the main sources of pollution; its outlet is in Bangreweogo urban park.
88 Dassasgho canal is 4.8 km long, it receives effluent from the city prison town and slaughterhouse as the main
89 sources of pollution; its outlet is in Bangreweogo. The geographical coordinates of the sampled canals were
90 Passpanga (12°23' N, 1°30' W), Zogona (12°23' N, 1°31' W), and Dassasgho (12°23'N, 1°29' W). From each

91 canal, samples were collected from the upper, middle, and lower reaches of the waterways. At each point, five
92 samples of 100 ml were collected around 2m², 15-20 cm below the surface, in sterile bottles. The five replicate
93 samples were combined to give one sample of 500 ml. The combined samples from each point were pooled
94 together to minimize the variations that may occur from one site to another, the representative sample of the
95 canal was 1500ml. The samples were transported on ice and frozen (-80°C) until processing. The values of
96 temperature, pH, total organic carbon and total nitrogen were for Passpanga (30°C, 8.27, 20.2 and 30.8,
97 respectively); Zogona (29.5°C, 7.20, 13.5 and 4.46, respectively) and Dassasgho (31.5°C, 7.39, 11.7 and 8.95,
98 respectively).

99 *DNA extraction and quantification*

100 To extract total genome DNA from the wastewater, samples (500 mL each) from each canal were
101 centrifuged at 2268 × g for 15 minutes to pellet the microbial cells. DNA extraction was conducted using the
102 FastDNA Soil Kit (MP Biomedicals, CA, USA), according to the manufacturer's instructions. Total extracted
103 DNA was quantified using the Quant-iT PicoGreen dsDNA Assay Kit, and the Qubit™ 3.0 Fluorometer (Qubit,
104 Life Technologies, USA).

105 *High-throughput sequencing*

106 DNA samples from each of the three canals were sent to Edinburgh Genomics (Edinburgh, Scotland)
107 for high-throughput sequencing using Illumina Hiseq4000 (Illumina, Inc, USA). Libraries were prepared by
108 TruSeq DNA Nano gel free library using 350 bp insert. The sequencing strategy was index PE150+8+150 cycle
109 (Paired End sequencing, 150-bp reads and 8-bp index sequence); it generated 76.3 Gb of raw sequence reads.
110 The metagenomic data have been deposited at National Center for Biotechnology Information (NCBI),
111 Sequence Read Archive (SRA) under project accession number PRJNA358310 (SRR5123280; SRR5123281;
112 SRR5123278; SRR5123291; SRR5123282; and SRR5123287).

113 *Quality reads filtering*

114 Raw fastq data were uploaded into GALAXY Cloudman, using the Cloud Infrastructure for Microbial
115 Bioinformatics (CLIMB) platform (Connor et al. 2016). Reads were assessed using FastQC (
116 <http://www.bioinformatics.babraham.ac.uk/projects/fastqc/>) (Andrews, 2014). Reads of bad quality were
117 filtered out by removing adapters, N bases below quality 3, a cutoff value of Phred score > 30, length sorted at ≥
118 70 bases, using Trimmomatic (Bolger et al. 2014). Passpanga, Zogona and Dassasgho had 83, 724, 220; 84, 013,
119 504 and 88, 861, 498 raw reads, for 64, 536, 398; 63, 920, 619 and 69, 695, 966 trimmed reads respectively. The

120 trimmed reads were submitted to annotation on MG-RAST metagenomics analysis server version 4.0.3 (Glass et
121 al. 2010). They were further processed for taxonomic, functional annotations, and plasmids identification.

122 *Taxonomic and functional annotations*

123 MetaPhlAn 2.0 (Metagenomic Phylogenetic Analysis) was used to profile the composition of microbial
124 communities with species level resolution in the datasets with 75% of nucleotide identity threshold, using
125 default parameter settings, and the sensitive local Bowtie2 alignment step. MetaPhlAn 2.0 program relies on ~1
126 million unique clade-specific marker genes identified from ~17,000 reference genomes (~13,500 bacterial and
127 archaeal, 3,500 viral, and 110 eukaryotic) (Segata et al., 2012). The additional analysis type and argument was
128 “-t_rel_ab_w_read_stats” to profile bacterial metagenomes to estimate the number of reads coming from each
129 clade. A cladogram was generated to visualize microbial abundances (circle size relating to microbial
130 abundance) on a tree of life using GraPhlAn (Asnicar et al. 2015). The functional classification of the
131 metagenomic data were annotated against SEED subsystems in MG-RAST (Glass et al. 2010). The functional
132 profiles were generated using the E-value cut-off 10^{-5} , a minimum identity of 80%, and a minimum alignment
133 length of 20 amino acids.

134 *Identification and quantification of antibiotic resistance and virulence factor gene markers.*

135 Antibiotic resistance genes (ARGs) and virulence factors (VFs) from metagenome reads were
136 identified and quantified using ShortBRED (Short Better Representative Extract Dataset) (Kaminski et al.
137 2015). ShortBRED profiles protein family abundance in metagenomes in two-steps: (i) *ShortBRED-Identify*
138 isolates representative peptide sequences (markers) for the protein families, and (ii) *ShortBRED-Quantify* maps
139 metagenomic reads against these markers to determine the relative abundance of their corresponding families.
140 Fragment length ≥ 30 amino acids and $\geq 95\%$ identity is considered a valid hit. ARG and VF markers were
141 generated from the Comprehensive Antibiotic Resistance Database (CARD) (McArthur et al. 2013) and the
142 Virulence Factor Database (VFDB) (Chen et al. 2005), respectively, and using UniRef50 as a reference protein
143 database. Using the antibiotic resistance ontology (ARO) numbers in CARD and VFDB, respectively, ARG and
144 VF markers were aggregated, annotated and associated to the corresponding resistance family or predicted
145 pathogen. Pathogen-specific and common VF genes were identified. Heatmaps were generated to visualize
146 tabular abundance of ARGs and VFs in the different canals with hclust2 (Asnicar et al. 2015), using Bray-Curtis
147 as the distance measure both between samples and between features.

148 *Identification of plasmid amplicons of clinical relevance*

149 Plasmid replicon sequences belonging to *Enterobacteriaceae* (Gram-negative) and to *Enterococcus*,
150 *Staphylococcus* and *Streptococcus* (Gram-positive) were downloaded from the PlasmidFinder database 1.3
151 (<https://cge.cbs.dtu.dk/services/PlasmidFinder>) (Carattoli et al. 2014), and aligned against the metagenomic
152 reads using BLAT. A read was assigned to a plasmid replicon if the BLAT hit had a sequence identity of $\geq 80\%$,
153 to detect both large and small plasmids (Carattoli et al. 2014), using the E-value cut-off 10^{-5} .

154

155 **RESULTS AND DISCUSSION**

156 *Microbial composition, population structure and functional classification*

157 Microbial population diversity was assessed to provide an overview on the diversity and relative
158 abundance of the different taxonomy groups, bacterial *phyla* and families of interest. The relative abundance of
159 bacteria, archaea and viruses is shown in Table 1. The relative abundance of *Bacteria* and *Archaea* domains of
160 Woese and viruses' group is shown in Table 1. The relative abundance of *Archaea* in samples from Passpanga
161 and Dassasgho were about 20 and 13-fold higher than the abundance in Zogona, respectively (Table S1). There
162 were 10 different bacteria *phyla* identified in Dassasgho, nine in Passpanga, and six in Zogona (Table 1).
163 *Proteobacteria* was the most prevalent *phylum* in all the wastewater samples tested, followed by *Actinobacteria*
164 and *Firmicutes* (Figure 2). Our results are in accordance with previous studies that showed that these *phyla* are
165 dominant in wastewaters (Mlejnkova and Sovova, 2010; Bengtsson-Palme et al. 2014). The microbial
166 community structure is significantly dependent on the level of water pollution, and abundance of *Proteobacteria*
167 or *Bacteroidetes* is reported to be a consequence of water pollution by human activities (Mlejnkova and Sovova,
168 2010; Suriya et al. 2017). Our data suggests that the sampled canals are impacted by anthropogenic pollution.
169 Natural water bodies such as rivers and lakes are less impacted and low nutrients content, thus favouring the
170 survival of autochthonous bacteria, which are mainly autotrophs (Sigee, 2004; Zeglin et al. 2015). The higher
171 bacterial diversity in the city canals could be explained by an increased nutrient load due to organic matter
172 pollution by anthropogenic activities, favouring the establishment of more allochthone bacteria communities
173 (Hewson et al. 2003; Qin et al. 2016). *Chlorobi*, *Fusobacteria*, *Spirochaetes* and *Verrucomicrobia* *phyla* have
174 members known to live in anaerobic or micro-anaerobic conditions (Nelson, 2015). Their presence in
175 wastewater could be a consequence of a decrease in oxygen availability following intense organic matter
176 degradation by microorganisms in the water bodies. *Fusobacteria*, *Spirochaetes*, and *Verrucomicrobia* are
177 commonly found in human and animal faeces, viruses are a major cause of waterborne diseases (Nelson, 2015;

178 Nieuwenhuijse et al. 2017). Their presence indicates that the water in these canals presents a high risk of
179 transmitting infectious diseases.

180 .

181 We assessed families of bacteria that carry contaminating plasmids ARGs and those that are
182 autochthonous bacteria in water ecosystems. The relative prevalence of the families *Enterobacteriaceae*,
183 *Enterococcaceae*, *Streptococcaceae*, *Staphylococcaceae*, *Comamonadaceae* and *Chroococcaceae* in the water
184 samples is shown in Table 1. Taking into consideration the wide standard deviation, the proportion of nitrate-
185 reducing bacteria (e.g. *Comamonadaceae*) and photosynthetic oxygenic microorganisms (e.g. *Chroococcaceae*)
186 was found to be similar to those bacteria containing antibiotic resistant plasmids. That indicates anthropogenic
187 pollution and ecological disturbances as suggested previously by others (Balmonte et al. 2016).

188 The functional categories of the microbial communities were distributed into 28 level 1 subsystems of
189 SEED database (Figure 3). The predominant functions were clustering-based subsystems (14.4%),
190 carbohydrates (13.4%), miscellaneous (10.8%), amino acids and derivatives (7.81%) and protein metabolism
191 (7%). Virulence, disease and defence accounted for 2.44%; potassium metabolism (0.28%), dormancy and
192 sporulation (0.20%) and photosynthesis (0.19%) were the less prevalent. Carbohydrates, amino acids and
193 derivatives, and protein metabolism were among the most prevalent subsystems in the city wastewaters. Our
194 results are in line with previous studies which have shown the predominance of those functions in wastewaters
195 (Wang et al. 2013; Chao et al. 2013; Bäumlisberger et al. 2015). Microbial communities residing in those waters
196 are well skilled to degrade easily accessible carbon substrates such as soluble carbohydrates or polysaccharides,
197 as well as amino acid derivatives and proteins (Uroz et al. 2013). This is consequent to a constant anthropogenic
198 pollution of these waters with organic matter. The proportion of genes involved in virulence, disease and
199 defence in the metagenome reads conformed with the range previously found in wastewater by others (Wang et
200 al. 2013). The negligible proportion of function allocated to photosynthesis translates the disappearance of the
201 bacterioplankton.

202 *Virulence factors*

203 To further assess the presence and relative abundance of bacteria pathogenic to humans in the
204 wastewaters, we assessed the presence of VF genes as pathogenic indicators. Sixty-one different VF genes were
205 identified in the samples (Figure 4). Eleven were pathogen-specific VFs and 60 are orthologous commonly
206 found in both pathogens and non-pathogens. The genes encoding different mechanisms of virulence were (in
207 order of relative abundance): secretion systems, toxins, exoenzymes, adherence, iron uptake, biofilm formation,

208 stress response, antiphagocytosis, manganese uptake, a type IV secretory protein, and plasminogen activator.
209 The identified VF genes are commonly carried by *Escherichia coli* (17), *Shigella* spp (14), *Clostridium*
210 *perfringens* (6) and *Mycobacterium tuberculosis* (6). The pathogen-specific VFs belonged to *Streptococcus*
211 *agalactiae*, *C. perfringens*, *M. tuberculosis*, *Legionella pneumophila*, *Shigella* spp, *S. flexneri*, *Yersinia*
212 *enterocolitica*, and *Bartonella henselae*. All the pathogen-specific VFs were present in Passpanga, those
213 belonging to *C. perfringens*, *S. flexneri*, *Y. enterocolitica*, and *B. henselae* were present in Dassasgho, and the
214 one of *Y. enterocolitica* in Zogona.

215 The widespread presence of VFs usually found in human pathogenic bacteria in wastewater using a
216 high-throughput shotgun sequencing technique has been reported in many countries (Lu et al. 2013; Ye and
217 Zhang, 2013; Kumaraswamy et al. 2014). However, in those studies the wastewater was neither from an open-
218 air canal nor used for irrigation of crops intended for human consumption. With regard to the number of
219 pathogen-specific and common VFs, next generation sequencing (NGS) is recommended to efficiently assess
220 the probable presence of pathogenic bacteria in wastewater. Its broader coverage allows the identification of
221 microorganisms present in low numbers, which would not be detected with traditional molecular tools
222 (Kumaraswamy et al. 2014). The presence of pathogen-specific and common VF genes in urban waterways and
223 which will be used for urban agriculture indicates that such waters represents a risk to human health. Most of
224 the detected VFs belong to pathogens that are transmitted by direct water exposure or ingestion of water/food
225 contaminated with human or animal faeces and are commonly responsible for waterborne diseases. Therefore, it
226 is not surprising that populations directly or indirectly exposed to these wastewaters could suffer from acute
227 diarrhoea, chronic gastritis, and gastroenteritis (Nitiema et al. 2011; Traore et al., 2015). In LMICs 842 000
228 people die annually from diarrhoea (WHO, 2017); this is because of inadequate water, sanitation, and hygiene.
229 Moreover, the presence of pathogen-specific VFs in these canals should be a source of concern, since the use of
230 this water for urban agriculture is a reservoir for pathogens and ARGs and so represents a strong vehicle for
231 transmission. Pathogen-specific VF belonging to *M. tuberculosis* was detected in Passpanga canal. *M.*
232 *tuberculosis* is predominantly transmitted between humans by droplets of water by coughing and possibly
233 sneezing. Its presence in the canal could be due to the high prevalence rate in the country; 5 866 cases of
234 tuberculosis were reported in Burkina Faso in 2015 (WHO, 2016). Patients infected with tuberculosis can
235 contaminate water and soil ecosystems, and *M. tuberculosis* can remain viable for extended periods of time
236 (Velayati et al. 2015). The indirect transmission of *M. tuberculosis* might be facilitated with their persistence in
237 the environment (Falkinham et al. 2015).

239 In total, 129 ARGs were detected; 81 were determined to be transmissible by mobile genetic elements
240 through horizontal gene transfer (HGT). The 81 transmissible ARGs confer resistance to 11 major classes of
241 antibiotics and trimethoprim (Figure 5). The most prevalent resistance genes encode resistance to
242 aminoglycosides (20 genes), tetracycline (19 genes), beta-lactams (13 genes) and macrolides (9 genes). ARG
243 prevalence in Zogona was the lowest, while prevalence of ARGs in Passpanga was the highest. A heat map of
244 the ARG relative abundance showed that Passpanga and Dassasgho clustered together; 69 ARGs were identified
245 in Passpanga, 21 in Zogona, and 55 in Dassasgho. Eighteen ARGs were common to the three sampling sites; 25
246 were common to Passpanga and Dassasgho; 25 and 10 resistant genes, respectively were found solely in
247 Passpanga and Dassasgho. Zogona had no unique ARG. The top five most prevalent transmissible ARGs were
248 *aadA15*, *aadA13*, *sul1*, *aadA*, and *tetB* in Passpanga; *sul1*, *sul2*, *aadA13*, *aadA*, and *RbpA* in Zogona; and *sul1*,
249 *sul2*, *aadA13*, *aadA*, and *aadA15* in Dassasgho. Interestingly, seven different ARGs encoding extended-
250 spectrum beta lactamases (ESBLs) (*bla*_{OXA-226}, *bla*_{OXA-256}, *bla*_{OXA-347}, *bla*_{OXA-46}, *bla*_{SHV-100}, *bla*_{GES-21}, and *bla*_{AIM-1})
251 were detected in the city wastewaters. All of these genes were present in Dassasgho. The *bla*_{SHV-100} and *bla*_{AIM-1}
252 genes were not found in Passpanga, but the *bla*_{OXA-226} and *bla*_{OXA-256} were identified in water from Zogona.

253 Water ecosystems polluted by pathogenic bacteria and ARGs act both as a natural reservoir and a
254 channel for the spread of clinically relevant antibiotic resistance traits. Previous studies have shown a
255 correlation between pathogen prevalence and ARGs (Michael et al. 2013; Bengtsson-Palme et al. 2014; Zheng et
256 al. 2017). Our study provides data that confirms those from previous studies, with greater relative prevalence of
257 pathogen-specific VFs and ARGs found in the canals. The wide variety of detected resistance gene classes
258 suggests that they might have been acquired under selection pressure from diverse antibiotics and/or other
259 factors selecting for resistances mechanisms (Kristiansson et al. 2011; Bengtsson-Palme et al. 2014). The high
260 density of bacteria in the canals, favours cell-to-cell contact, which likely contributes to ARG transmission
261 (Tennstedt et al. 2003; Bengtsson-Palme et al. 2014). Hospital wastewater is known to contain high quantities of
262 drug-resistant bacteria and antibiotic residues, and their interaction contributes to an increase of ARGs (Brecht
263 et al., 2014; Lien et al. 2016). This could explain the highest ARGs diversity and relative abundance in
264 Passpanga, which receives effluent from the university teaching hospital.

265 Aminoglycosides, sulphonamides, tetracyclines and beta-lactams are the most commonly administered
266 antibiotics in human medicine in Burkina Faso (Krause et al. 1999; Ouedraogo et al. 2017). Tetracyclines and
267 sulphonamides are commonly used as growth promoters and for disease prevention in animals (Zhang et al.

268 2015). This could explain the high prevalence of genes encoding resistance against these antibiotics in the
269 canals. Aminoglycoside-modifying enzymes (*aac*, *aad*, and *ant*) are plasmid and transposon associated;
270 dihydropteroate synthase enzymes (*sul1* and *sul2*) are plasmid encoded and found on class 1 integrons. These
271 characteristics facilitate the dissemination and persistence of drug resistance phenotypes within the bacterial
272 communities, and thus could explain their relative predominance in the canal water (Hall and Collis, 1995;
273 Mingeot-Leclercq et al. 1999; Stoll et al., 2012). The relatively high prevalence and diversity of tetracycline
274 resistance genes identified in our datasets could be related to their presence on mobile genetic elements and
275 single strains carrying multiple drug-resistance genes. It was of concern to find seven different ESBL genes
276 encoding synthesis of four enzyme families in these ecosystems. It was reported that 58% of *Enterobacteriaceae*
277 responsible for infections in three main hospitals of Burkina Faso were ESBL producers (Ouedraogo et al.,
278 2017). Our results suggest a link between resistance in clinical settings and in the environment like previously
279 stated (Tacao et al. 2012; Forsberg et al. 2012). Two genes (*bla*_{AIM-1} and *bla*_{GES-21}) encoding carbapenemase
280 enzymes were found in the wastewater. Carbapenems are the last resort drugs for patients infected with ESBL-
281 producing bacteria; they are expensive drugs, which many patients in LMICs cannot afford.

282 *Plasmid replicons*

283 The presence of plasmids, putative carriers and disseminative agents of ARGs by horizontal gene
284 transfer was assessed. Thirty-one plasmid replicons of medical interest from *Enterobacteriaceae* were identified
285 in the wastewater; 27 were present in Passpanga, four in Zogona and 18 in Dassasgho (Table S2). *E. coli* was
286 the main likely carrier. The identified plasmid replicons belonged to 10 different incompatibility groups, namely
287 ColE, FIB, FIC, FII, P, Q, R, U, Y, and A/C. The 10 groups were found in Passpanga and Dassasgho, while
288 only IncP and IncQ were identified in Zogona. The three most prevalent plasmid replicon families were IncFIB,
289 IncFII and IncQ in Passpanga, and IncQ, ColE and IncP in Dassasgho. Thirty plasmid replicon types from
290 Gram-positive bacteria were found in the canal water; all were present in Passpanga, 21 in Dassasgho, and none
291 in Zogona (Table S.3). In both the Passpanga and Dassasgho canals, based on the probe origin, *Staphylococcus*
292 spp was the most likely carrier, followed by *Enterococcus* spp. One plasmid replicon type in Passpanga was
293 associated with *Streptococcus* spp as a possible carrier.

294 Plasmids can carry genes conferring resistance to one or more antibiotics, and their ability to transfer
295 by HGT between bacteria make them key contributors to the emergence and dissemination of ARGs in human
296 and animal medicine (Orlek et al. 2017). The great relative number of plasmid replicons in the canals is a
297 consequence to plasmid production and acquisition (Wegrzyn et al. 2002; Kristiansson et al. 2011). The

298 diversity of plasmid incompatibility group and relative abundance in the different canals matched those of
299 ARGs, suggesting likely association between the two. The identified plasmid incompatibility groups are
300 carriers of ARGs encoding resistance to all common antibiotic classes. Mobile genetic elements are known to
301 greatly contribute to the dissemination of ARGs via HGT and plasmids are responsible for much of the
302 widespread distribution of clinically relevant ARGs (Gyles and Boerlin, 2014). The presence of such extensive
303 plasmid incompatibility groups could contribute to their extensive acquisition by different bacterial species, and
304 further spread. The plasmid replicons IncP and Q are frequently associated with the carbapenemases *bla*_{AIM-1}
305 and *bla*_{GES-21}, respectively (Carattoli, 2009; Leiros et al. 2012); all were found in both Passpanga and
306 Dassasgho. *E. coli* is a faecal indicator, its presence indicates the possible occurrence of pathogenic bacteria
307 and faecal contamination. *E. coli* was the main carrier of *Enterobacteriaceae* plasmid replicons of medical
308 interest in the waters. Considering its plasticity, *E. coli* easily acquires and transfer ARGs to other strains and
309 intensively promotes bacterial resistance dissemination (Rasheed et al. 2014). Thus, these wastewaters
310 constitute a serious health risk. The multidrug resistant pathogens of special concern for both community
311 and/or hospital acquired infections are *E. coli*, *K. pneumoniae*, *S. aureus*, *S. pneumoniae*, *Enterococci*, *V.*
312 *cholerae*, and *M. tuberculosis*. Pathogen-specific and common VFs were found in the wastewaters along with
313 clinically relevant plasmids and ARGs conferring resistance to all major classes of antibiotics. Consequently,
314 these wastewaters can give rise to drug-resistant bacteria and so represent a potentially serious threat to public
315 health in the town via direct contact or the food chain. It is estimated that vegetable production occupies more
316 than 561 ha in the city, using part of the 620,000m³ of both domestic and industrial wastewater that are
317 discharged every year (Vezina, 2002; Dao et al. 2016).

318

319 **Conclusion**

320 Metagenomic exploration of wastewater in open-air canals, used for urban agriculture in Ouagadougou,
321 revealed a wide functional specialization of microbial communities, a high prevalence of pathogen-specific and
322 commonly found VFs, a variety of plasmid incompatibility types and multiple ARGs in this environment,
323 suggesting that wastewater is a “hot spot” for formation and dissemination of antibiotic resistance. Our data
324 showed that wastewater for urban agriculture in Burkina Faso might contribute to dissemination of bacterial
325 resistance, including ESBLs and carbapenemases. That is of special concern for both community and hospital
326 acquired infections. Further investigations are needed to determine the extent that exposed populations are

327 affected by this health issue. There is an urgent need to improve global access to clean water, sanitation, and
328 hygiene in LMICs to prevent the rise and dissemination of bacterial resistance from the environment to people.

329

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336

337 **Transparency declarations**

338 The authors declare no competing financial interest.

339

340 **References**

341

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