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1	Wastewater used for urban agriculture in West Africa as a reservoir for
2	antibacterial resistance dissemination
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12	
13	Summary
14	State of art metagenomics were used to investigate the microbial population, antibiotic resistance genes and
15	plasmids of medical interest in wastewater used for urban agriculture in Ouagadougou (Burkina Faso).
16	Wastewater samples were collected from three canals near agricultural fields in three neighbourhoods.
17	Assessment of microbial population diversity revealed different microbial patterns among the different samples.
18	Sequencing reads from the wastewaters revealed different functional specializations of microbial communities,
19	with the predominance of carbohydrates and proteins metabolism functions. Eleven pathogen-specific and 56
20	orthologous virulence factor genes were detected in the wastewater samples. These virulence factors are usually
21	found in human pathogens that cause gastroenteritis and/or diarrhoea. A wide range of antibiotic resistance
22	genes was identified; 81 are transmissible by mobile genetic elements. These included seven different extended
23	spectrum β -lactamase genes encoding synthesis of four enzyme families, including two metallo- β -lactamases
24	(bla _{AIM-1} and bla _{GES-21}). Ten different incompatibility groups of Enterobacteriaceae plasmid replicons (ColE,
25	FIB, FIC, FII, P, Q, R, U, Y, and A/C), and 30 plasmid replicon types from Gram-positive bacteria. All are
26	implicated in the wide distribution of antibiotic resistance genes. We conclude that wastewater used for urban
27	agriculture in the city represents a high risk for spreading bacteria and antimicrobial resistance among humans
28	and animals.
29	Keywords: Antimicrobial resistance, metagenomics, urban agriculture, low and middle-income countries

31 Introduction

32 There is an increasing number of bacteria that are multi-resistant against common antibiotics and thus 33 cannot be treated by current therapies (Leopold et al. 2014). Antibiotic resistance has led to the need for more 34 expensive drugs, which many cannot afford, resulting in increased morbidity and mortality (Laxminarayan et al. 35 2016). Selection of drug-resistant bacteria and proliferation occurs at sub-inhibitory concentrations of antibiotics 36 (Drlica, 2003; Andersson and Hughes, 2010). It has been reported that approximately 50 - 90 % of antibiotics 37 administered to humans and animals are excreted via urine and faeces, as a mixture of parent drug and 38 metabolite forms, and thus significant levels of active drug end up in the environment, where they may persist in 39 soil and aquatic ecosystems (Kümmerer, 2009). To combat the presence of antibiotics in ecosystems, bacteria 40 have evolved a plethora of different resistance genes of which many are mobile and can easily spread between 41 species including human pathogens (Groh et al. 2007; Wellington et al. 2013; Zhang et al. 2016). Consequently, 42 environmental bacteria serve as a reservoir of resistance genes that can be transmitted to pathogenic species 43 (Allen et al. 2009). That is of concern in low and middle-income countries (LMICs) where populations are often 44 directly exposed to untreated wastewater due to a lack of water and sanitation services including in urban areas 45 (UN, 2017; Bougnom and Piddock, 2017).

46 The urban population in sub-Saharan Africa is projected to grow by 3.5 % per annum from now until 47 2030, and the number of urban residents is expected to rise from 400 million in 2010 to 1.26 billion in 2050 48 (UN, 2012). This rapid increase in the urban population raises significant challenges, related to urban poverty, 49 public health, housing, food security and environmental pollution. Urban and peri-urban agriculture is defined 50 as the production of crops and livestock within and around cities (UNDP, 1996). It has been developed by urban 51 dwellers to meet their food demand, as well as a source of employment and income. The importance of urban 52 agriculture in terms of contributing to food security and poverty alleviation has been recognised and promoted 53 (Martellozzo et al. 2014). Two hundred million urban dwellers are reported to be engaged in urban agriculture 54 worldwide, producing in some cases up to 90 % of cities' demand for perishable vegetables (UN, 2013). It is 55 reported that more than 80% of the domestic and industrial wastewater generated in LMICs is discharged 56 untreated into the environment, and because of its low cost, availability, and nutrient content, urban agriculture 57 hugely relies on wastewater for irrigation, (Mateo-Sagasta et al. 2013). As water is a vehicle for microbial 58 dissemination, wastewater for agricultural irrigation represents a very serious health risk, not least as it increases 59 exposure to faecal pathogens. Ingestion of faecally-contaminated water and/or food by opportunistic and/or 60 pathogenic microorganisms is one of the major reasons for the high number of water-related diseases in LMICs.

61 Wastewater use for urban agriculture has been shown to expose humans and animals to enteric diseases caused 62 by pathogenic bacteria, protozoa, and helminths (Dickin et al. 2016). To reduce the spread of antimicrobial 63 resistant bacteria to humans, the role of wastewater, sanitation and hygiene must be understood and addressed 64 (Graham et al. 2011; Wuijts et al. 2017).

Genomics allows us to understand the factors that drive ARGs transfer between bacteria and the mechanisms conferring resistance (Amos et al. 2014). The use of a metagenomic approach provides an opportunity to provide essential information so that the risks of untreated wastewater and impact upon human and animal health can be defined. In this study, we used metagenomics to assess the presence of virulence factor genes (VFs), antibiotic resistance genes (ARGs) and plasmids of medical interest in wastewater used for urban agriculture in Ouagadoudou (Burkina Faso). The aim was to measure the potential of urban agriculture to disseminate antibiotic resistance in the city.

72

73 Material and Methods

74 Wastewater sampling

75 Experiments were conducted with samples collected from Ouagadougou (Burkina Faso, West Africa) 76 in October 2015, at the end of the rainy season, with three precipitation events during the month. Ouagadougou 77 is the capital city of Burkina Faso, it occupies a surface area of 219.3 km², with a population of about 2.7 78 million inhabitants. The city is in the Sudano-Sahelian area, with a rainfall of about 800 mm per year, and 79 temperatures ranging from 16 °C to 43 °C during the rainy and dry seasons, respectively. Wastewater samples 80 used for urban agriculture were collected in three neighborhoods from three canals near agricultural fields. The 81 sampling sites were selected based on their different characteristics. These canals are open-air water drainage 82 and collection points of different transects (Figure 1). They were constructed to protect the city from floods, and 83 receive hand-draining sludge, solid waste, and wastewater. They do not have regular cleaning and are used to 84 irrigate agricultural fields which are found along its way. Passpanga canal is 5 km long, it receives effluent from 85 the university teaching hospital, the city and biggest market of the town; these are the main sources of pollution. 86 Its outlet is dam N° 3. Zogona canal, which is 4.4 km long; it receives effluent from 1200 houses and the 87 University of Ouagadougou as the main sources of pollution; its outlet is in Bangreweogo urban park. 88 Dassasgho canal is 4.8 km long, it receives effluent from the city prison town and slaughterhouse as the main 89 sources of pollution; its outlet is in Bangreweogo. The geographical coordinates of the sampled canals were 90 Passpanga (12°23' N, 1°30' W), Zogona (12°23' N, 1°31' W), and Dassasgho (12°23'N, 1°29' W). From each

91 canal, samples were collected from the upper, middle, and lower reaches of the waterways. At each point, five 92 samples of 100 ml were collected around $2m^2$, 15-20 cm below the surface, in sterile bottles. The five replicate 93 samples were combined to give one sample of 500 ml. The combined samples from each point were pooled 94 together to minimize the variations that may occur from one site to another, the representative sample of the 95 canal was1500ml. The samples were transported on ice and frozen (-80°C) until processing. The values of 96 temperature, pH, total organic carbon and total nitrogen were for Passpanga (30°C, 8.27, 20.2 and 30.8, 97 respectively); Zogona (29.5°C, 7.20, 13.5 and 4.46, respectively) and Dassasgho (31.5°C, 7.39, 11.7 and 8.95, 98 respectively).

99 DNA extraction and quantification

100 To extract total genome DNA from the wastewater, samples (500 mL each) from each canal were 101 centrifuged at $2268 \times g$ for 15 minutes to pellet the microbial cells. DNA extraction was conducted using the 102 FastDNA Soil Kit (MP Biomedicals, CA, USA), according to the manufacturer's instructions. Total extracted 103 DNA was quantified using the Quant-iT PicoGreen dsDNA Assay Kit, and the QubitTM 3.0 Fluorometer (Qubit, 104 Life Technologies, USA).

105 High-throughput sequencing

106DNA samples from each of the three canals were sent to Edinburgh Genomics (Edinburgh, Scotland)107for high-throughput sequencing using Illumina Hiseq4000 (Illumina, Inc, USA). Libraries were prepared by108TruSeq DNA Nano gel free library using 350 bp insert. The sequencing strategy was index PE150+8+150 cycle109(Paired End sequencing, 150-bp reads and 8-bp index sequence); it generated 76.3 Gb of raw sequence reads.110The metagenomic data have been deposited at National Center for Biotechnology Information (NCBI),111Sequence Read Archive (SRA) under project accession number PRJNA358310 (SRR5123280; SRR5123281;112SRR5123278; SRR5123291; SRR5123282; and SRR5123287).

113 *Quality reads filtering*

114Raw fastq data were uploaded into GALAXY Cloudman, using the Cloud Infrastructure for Microbial115Bioinformatics (CLIMB) platform (Connor et al. 2016). Reads were assessed using FastQC (116http://www.bioinformatics.babraham.ac.uk/projects/fastqc/)(117filtered out by removing adapters, N bases below quality 3, a cutoff value of Phred score > 30, length sorted at \geq 11870 bases, using Trimmomatic (Bolger et al. 2014). Passpanga, Zogona and Dassasgho had 83, 724, 220; 84, 013,119504 and 88, 861, 498 raw reads, for 64, 536, 398; 63, 920, 619 and 69, 695, 966 trimmed reads respectively. The

120 trimmed reads were submitted to annotation on MG-RAST metagenomics analysis server version 4.0.3 (Glass et

121 al. 2010). They were further processed for taxonomic, functional annotations, and plasmids identification.

122 Taxonomic and functional annotations

123 MetaPhlAn 2.0 (Metagenomic Phylogenetic Analysis) was used to profile the composition of microbial 124 communities with species level resolution in the datasets with 75% of nucleotide identity threshold, using 125 default parameter settings, and the sensitive local Bowtie2 alignment step. MetaPhlAn 2.0 program relies on ~1 126 million unique clade-specific marker genes identified from ~17,000 reference genomes (~13,500 bacterial and 127 archaeal, 3,500 viral, and 110 eukaryotic) (Segata et al., 2012). The additional analysis type and argument was 128 "-t rel_ab_w_read_stats" to profile bacterial metagenomes to estimate the number of reads coming from each 129 clade. A cladogram was generated to visualize microbial abundances (circle size relating to microbial 130 abundance) on a tree of life using GraPhlAn (Asnicar et al. 2015). The functional classification of the 131 metagenomic data were annotated against SEED subsystems in MG-RAST (Glass et al. 2010). The functional 132 profiles were generated using the E-value cut-off 10^{-5} , a minimum identity of 80%, and a minimum alignment 133 length of 20 amino acids.

134 Identification and quantification of antibiotic resistance and virulence factor gene markers.

135 Antibiotic resistance genes (ARGs) and virulence factors (VFs) from metagenome reads were 136 identified and quantified using ShortBRED (Short Better Representative Extract Dataset) (Kaminski et al. 137 2015). ShortBRED profiles protein family abundance in metagenomes in two-steps: (i) ShortBRED-Identify 138 isolates representative peptide sequences (markers) for the protein families, and (ii) ShortBRED-Quantify maps 139 metagenomic reads against these markers to determine the relative abundance of their corresponding families. 140 Fragment length \geq 30 amino acids and \geq 95% identity is considered a valid hit. ARG and VF markers were 141 generated from the Comprehensive Antibiotic Resistance Database (CARD) (McArthur et al. 2013) and the 142 Virulence Factor Database (VFDB) (Chen et al. 2005), respectively, and using UniRef50 as a reference protein 143 database. Using the antibiotic resistance ontology (ARO) numbers in CARD and VFDB, respectively, ARG and 144 VF markers were aggregated, annotated and associated to the corresponding resistance family or predicted 145 pathogen. Pathogen-specific and common VF genes were identified. Heatmaps were generated to visualize 146 tabular abundance of ARGs and VFs in the different canals with hclust2 (Asnicar et al. 2015), using Bray-Curtis 147 as the distance measure both between samples and between features.

148 Identification of plasmid amplicons of clinical relevance

Plasmid replicon sequences belonging to *Enterobacteriaceae* (Gram-negative) and to *Enterococcus*, Staphylococcus and Streptococcus (Gram-positive) were downloaded from the PlasmidFinder database 1.3 (<u>https://cge.cbs.dtu.dk/services/PlasmidFinder</u>) (Carattoli et al. 2014), and aligned against the metagenomic reads using BLAT. A read was assigned to a plasmid replicon if the BLAT hit had a sequence identity of \geq 80%,

153 to detect both large and small plasmids (Carattoli et al. 2014), using the E-value cut-off 10^{-5} .

154

155 RESULTS AND DISCUSSION

156

Microbial composition, population structure and functional classification

157 Microbial population diversity was assessed to provide an overview on the diversity and relative 158 abundance of the different taxonomy groups, bacterial phyla and families of interest. The relative abundance of 159 bacteria, archaea and viruses is shown in Table 1. The relative abundance of Bacteria and Archaea domains of 160 Woese and viruses' group is shown in Table 1. The relative abundance of Archaea in samples from Passpanga 161 and Dassasgho were about 20 and 13-fold higher than the abundance in Zogona, respectively (Table S1). There 162 were 10 different bacteria phyla identified in Dassasgho, nine in Passpanga, and six in Zogona (Table 1). 163 Proteobacteria was the most prevalent phylum in all the wastewater samples tested, followed by Actinobacteria 164 and *Firmicutes* (Figure 2). Our results are in accordance with previous studies that showed that these *phyla* are 165 dominant in wastewaters (Mlejnkova and Sovova, 2010; Bengtsson-Palme et al. 2014). The microbial 166 community structure is significantly dependent on the level of water pollution, and abundance of *Proteobacteria* 167 or Bacteroidetes is reported to be a consequence of water pollution by human activities (Mlejnkova and Sovova, 168 2010; Suriya et al. 2017). Our data suggests that the sampled canals are impacted by anthropogenic pollution. 169 Natural water bodies such as rivers and lakes are less impacted and low nutrients content, thus favouring the 170 survival of autochthonous bacteria, which are mainly autotrophs (Sigee, 2004; Zeglin et al. 2015). The higher 171 bacterial diversity in the city canals could be explained by an increased nutrient load due to organic matter 172 pollution by anthropogenic activities, favouring the establishment of more allochthone bacteria communities 173 (Hewson et al. 2003; Qin et al. 2016). Chlorobi, Fusobacteria, Spirochaetes and Verrucomicrobia phyla have 174 members known to live in anaerobic or micro-anaerobic conditions (Nelson, 2015). Their presence in 175 wastewater could be a consequence of a decrease in oxygen availability following intense organic matter 176 degradation by microorganisms in the water bodies. Fusobacteria, Spirochaetes, and Verrucomicrobia are 177 commonly found in human and animal faeces, viruses are a major cause of waterborne diseases (Nelson, 2015; 178 Nieuwenhuijse et al. 2017). Their presence indicates that the water in these canals presents a high risk of 179 transmitting infectious diseases.

180

We assessed families of bacteria that carry contaminating plasmids ARGS and those that are autochthonous bacteria in water ecosystems. The relative prevalence of the families *Enterobacteriaceae*, *Enterococcaceae*, *Streptococcaceae*, *Staphylococcaceae*, *Comamonadaceae* and *Chroococcaceae* in the water samples is shown in Table 1. Taking into consideration the wide standard deviation, the proportion of nitratereducing bacteria (e.g. *Comamonadaceae*) and photosynthetic oxygenic microorganisms (e.g. *Chroococcaceae*) was found to be similar to those bacteria containing antibiotic resistant plasmids. That indicates anthropogenic pollution and ecological disturbances as suggested previously by others (Balmonte et al. 2016).

188 The functional categories of the microbial communities were distributed into 28 level 1 subsystems of 189 SEED database (Figure 3). The predominant functions were clustering-based subsystems (14.4%), 190 carbohydrates (13.4%), miscellaneous (10.8%), amino acids and derivatives (7.81%) and protein metabolism 191 (7%). Virulence, disease and defence accounted for 2.44%; potassium metabolism (0.28%), dormancy and 192 sporulation (0.20%) and photosynthesis (0.19%) were the less prevalent. Carbohydrates, amino acids and 193 derivatives, and protein metabolism were among the most prevalent subsystems in the city wastewaters. Our 194 results are in line with previous studies which have shown the predominance of those functions in wastewaters 195 (Wang et al. 2013; Chao et al. 2013; Bäumlisberger et al. 2015). Microbial communities residing in those waters 196 are well skilled to degrade easily accessible carbon substrates such as soluble carbohydrates or polysaccharides, 197 as well as amino acid derivatives and proteins (Uroz et al. 2013). This is consequent to a constant anthropogenic 198 pollution of these waters with organic matter. The proportion of genes involved in virulence, disease and 199 defence in the metagenome reads conformed with the range previously found in wastewater by others (Wang et 200 al. 2013). The negligible proportion of function allocated to photosynthesis translates the disappearance of the 201 bacterioplankton.

202 Virulence factors

To further assess the presence and relative abundance of bacteria pathogenic to humans in the wastewaters, we assessed the presence of VF genes as pathogenic indicators. Sixty-one different VF genes were identified in the samples (Figure 4). Eleven were pathogen-specific VFs and 60 are orthologous commonly found in both pathogens and non-pathogens. The genes encoding different mechanisms of virulence were (in order of relative abundance): secretion systems, toxins, exoenzymes, adherence, iron uptake, biofilm formation, stress response, antiphagocytosis, manganese uptake, a type IV secretory protein, and plasminogen activator.
The identified VF genes are commonly carried by *Escherichia coli* (17), *Shigella* spp (14), *Clostridium perfringens* (6) and *Mycobacterium tuberculosis* (6). The pathogen-specific VFs belonged to *Streptococcus agalactiae*, C. *perfringens*, M. *tuberculosis*, *Legionnela pneumophila*, *Shigella* spp, S. *flexneri*, *Yersinia enterocolitica*, and *Bartonnella henselae*. All the pathogen-specific VFs were present in Passpanga, those belonging to C. *perfringens*, S. *flexneri*, Y. *enterocolitica*, and B. *henselae* were present in Dassasgho, and the one of Y. *enterocolitica* in Zogona.

215 The widespread presence of VFs usually found in human pathogenic bacteria in wastewater using a 216 high-throughput shotgun sequencing technique has been reported in many countries (Lu et al. 2013; Ye and 217 Zhang, 2013; Kumaraswamy et al. 2014). However, in those studies the wastewater was neither from an open-218 air canal nor used for irrigation of crops intended for human consumption. With regard to the number of 219 pathogen-specific and common VFs, next generation sequencing (NGS) is recommended to efficiently assess 220 the probable presence of pathogenic bacteria in wastewater. Its broader coverage allows the identification of 221 microorganisms present in low numbers, which would not be detected with traditional molecular tools 222 (Kumaraswamy et al. 2014). The presence of pathogen-specific and common VF genes in urban waterways and 223 which will be used for urban agriculture indicates that such waters represents a risk to human health. Most of 224 the detected VFs belong to pathogens that are transmitted by direct water exposure or ingestion of water/food 225 contaminated with human or animal faeces and are commonly responsible for waterborne diseases. Therefore, it 226 is not surprising that populations directly or indirectly exposed to these wastewaters could suffer from acute 227 diarrhoea, chronic gastritis, and gastroenteritis (Nitiema et al. 2011; Traore et al., 2015). In LMICs 842 000 228 people die annually from diarrhoea (WHO, 2017); this is because of inadequate water, sanitation, and hygiene. 229 Moreover, the presence of pathogen-specific VFs in these canals should be a source of concern, since the use of 230 this water for urban agriculture is a reservoir for pathogens and ARGs and so represents a strong vehicle for 231 transmission. Pathogen-specific VF belonging to M. tuberculosis was detected in Passpanga canal. M. 232 tuberculosis is predominantly transmitted between humans by droplets of water by coughing and possibly 233 sneezing. Its presence in the canal could be due to the high prevalence rate in the country; 5 866 cases of 234 tuberculosis were reported in Burkina Faso in 2015 (WHO, 2016). Patients infected with tuberculosis can 235 contaminate water and soil ecosystems, and M. tuberculosis can remain viable for extended periods of time 236 (Velayati et al. 2015). The indirect transmission of M. tuberculosis might be facilitated with their persistence in 237 the environment (Falkinham et al. 2015).

238 Antibiotic resistance genes analysis

239 In total, 129 ARGs were detected; 81 were determined to be transmissible by mobile genetic elements 240 through horizontal gene transfer (HGT). The 81 transmissible ARGs confer resistance to 11 major classes of 241 antibiotics and trimethoprim (Figure 5). The most prevalent resistance genes encode resistance to 242 aminoglycosides (20 genes), tetracycline (19 genes), beta-lactams (13 genes) and macrolides (9 genes). ARG 243 prevalence in Zogona was the lowest, while prevalence of ARGs in Passpanga was the highest. A heat map of 244 the ARG relative abundance showed that Passpanga and Dassasgho clustered together; 69 ARGs were identified 245 in Passpanga, 21 in Zogona, and 55 in Dassasgho. Eighteen ARGs were common to the three sampling sites; 25 246 were common to Passpanga and Dassasgho; 25 and 10 resistant genes, respectively were found solely in 247 Passpanga and Dassasgho. Zogona had no unique ARG. The top five most prevalent transmissible ARGs were 248 aadA15, aadA13, sul1, aadA, and tetB in Passpanga; sul1, sul2, aadA13, aadA, and RbpA in Zogona; and sul1, 249 sul2, aadA13, aadA, and aadA15 in Dassasgho. Interestingly, seven different ARGs encoding extended-250 spectrum beta lactamases (ESBLs) (bla_{OXA-226}, bla_{OXA-256}, bla_{OXA-347}, bla_{OXA-46}, bla_{SHV-100}, bla_{GES-21}, and bla_{AIM-1}) 251 were detected in the city wastewaters. All of these genes were present in Dassasgho. The bla_{SHV-100} and bla_{AIM-1} 252 genes were not found in Passpanga, but the $bla_{OXA-226}$ and $bla_{OXA-256}$ were identified in water from Zogona.

253 Water ecosystems polluted by pathogenic bacteria and ARGs act both as a natural reservoir and a 254 channel for the spread of clinically relevant antibiotic resistance traits. Previous studies have shown a 255 correlation between pathogen prevalence and ARGs (Michael et al. 2013; Bengtsson-Palme et al. 2014; Zheng et 256 al. 2017). Our study provides data that confirms those from previous studies, with greater relative prevalence of 257 pathogen-specific VFs and ARGs found in the canals. The wide variety of detected resistance gene classes 258 suggests that they might have been acquired under selection pressure from diverse antibiotics and/or other 259 factors selecting for resistances mechanisms (Kristiansson et al. 2011; Bengtsson-Palme et al. 2014). The high 260 density of bacteria in the canals, favours cell-to-cell contact, which likely contributes to ARG transmission 261 (Tennstedt et al. 2003; Bengtsson-Palme et al. 2014). Hospital wastewater is known to contain high quantities of 262 drug-resistant bacteria and antibiotic residues, and their interaction contributes to an increase of ARGs (Brechet 263 et al., 2014; Lien et al. 2016). This could explain the highest ARGs diversity and relative abundance in 264 Passpanga, which receives effluent from the university teaching hospital.

Aminoglycosides, sulphonamides, tetracyclines and beta-lactams are the most commonly administered antibiotics in human medicine in Burkina Faso (Krause et al. 1999; Ouedraogo et al. 2017). Tetracyclines and sulphonamides are commonly used as growth promoters and for disease prevention in animals (Zhang et al. 268 2015). This could explain the high prevalence of genes encoding resistance against these antibiotics in the 269 canals. Aminoglycoside-modifying enzymes (aac, aad, and ant) are plasmid and transposon associated; 270 dihydropteroate synthase enzymes (sul1 and sul2) are plasmid encoded and found on class 1 integrons. These 271 characteristics facilitate the dissemination and persistence of drug resistance phenotypes within the bacterial 272 communities, and thus could explain their relative predominance in the canal water (Hall and Collis, 1995; 273 Mingeot-Leclercq et al. 1999; Stoll et al., 2012). The relatively high prevalence and diversity of tetracycline 274 resistance genes identified in our datasets could be related to their presence on mobile genetic elements and 275 single strains carrying multiple drug-resistance genes. It was of concern to find seven different ESBL genes 276 encoding synthesis of four enzyme families in these ecosystems. It was reported that 58% of Enterobactericeae 277 responsible for infections in three main hospitals of Burkina Faso were ESBL producers (Ouedraogo et al., 278 2017). Our results suggest a link between resistance in clinical settings and in the environment like previously 279 stated (Tacao et al. 2012; Forsberg et al. 2012). Two genes (*bla*_{AIM-1} and *bla*_{GES-21}) encoding carbapenemase 280 enzymes were found in the wastewater. Carbapenems are the last resort drugs for patients infected with ESBL-281 producing bacteria; they are expensive drugs, which many patients in LMICs cannot afford.

282 Plasmid replicons

283 The presence of plasmids, putative carriers and disseminative agents of ARGs by horizontal gene 284 transfer was assessed. Thirty-one plasmid replicons of medical interest from Enterobacteriaceae were identified 285 in the wastewater; 27 were present in Passpanga, four in Zogona and 18 in Dassasgho (Table S2). E. coli was 286 the main likely carrier. The identified plasmid replicons belonged to 10 different incompatibility groups, namely 287 ColE, FIB, FIC, FII, P, Q, R, U, Y, and A/C. The 10 groups were found in Passpanga and Dassasgho, while 288 only IncP and IncQwere identified in Zogona. The three most prevalent plasmid replicon families were IncFIB, 289 IncFII and IncQ in Passpanga, and IncQ, ColE and IncP in Dassasgho. Thirty plasmid replicon types from 290 Gram-positive bacteria were found in the canal water; all were present in Passpanga, 21 in Dassasgho, and none 291 in Zogona (Table S.3). In both the Passpanga and Dassasgho canals, based on the probe origin, Staphylococcus 292 spp was the most likely carrier, followed by *Enterococcus* spp. One plasmid replicon type in Passpanga was 293 associated with Streptoccoccus spp as a possible carrier.

Plasmids can carry genes conferring resistance to one or more antibiotics, and their ability to transfer by HGT between bacteria make them key contributors to the emergence and dissemination of ARGs in human and animal medicine (Orlek et al. 2017). The great relative number of plasmid replicons in the canals is a consequence to plasmid production and acquisition (Wegrzyn et al. 2002; Kristiansson et al. 2011). The 298 diversity of plasmid incompatibility group and relative abundance in the different canals matched those of 299 ARGs, suggesting likely association between the two. The identified plasmid incompatibility groups are 300 carriers of ARGs encoding resistance to all common antibiotic classes. Mobile genetic elements are known to 301 greatly contribute to the dissemination of ARGs via HGT and plasmids are responsible for much of the 302 widespread distribution of clinically relevant ARGs (Gyles and Boerlin, 2014). The presence of such extensive 303 plasmid incompatibility groups could contribute to their extensive acquisition by different bacterial species, and 304 further spread. The plasmid replicons IncP and Q are frequently associated with the carbapenemases bla_{AIM-1} 305 and $bla_{GFS,21}$, respectively (Carattoli, 2009; Leiros et al. 2012); all were found in both Passpanga and 306 Dassasgho. E. coli is a faecal indicator, its presence indicates the possible occurrence of pathogenic bacteria 307 and faecal contamination. E. coli was the main carrier of Enterobacteriaceae plasmid replicons of medical 308 interest in the waters. Considering its plasticity, E. coli easily acquires and transfer ARGs to other strains and 309 intensively promotes bacterial resistance dissemination (Rasheed et al. 2014). Thus, these wastewaters 310 constitute a serious health risk. The multidrug resistant pathogens of special concern for both community 311 and/or hospital acquired infections are E. coli, K. pneumoniae, S. aureus, S. pneumoniae, Enterococci, V. 312 cholerae, and M. tuberculosis. Pathogen-specific and common VFs were found in the wastewaters along with 313 clinically relevant plasmids and ARGs conferring resistance to all major classes of antibiotics. Consequently, 314 these wastewaters can give rise to drug-resistant bacteria and so represent a potentially serious threat to public 315 health in the town via direct contact or the food chain. It is estimated that vegetable production occupies more 316 than 561 ha in the city, using part of the $620,000m^3$ of both domestic and industrial wastewater that are 317 discharged every year (Vezina, 2002; Dao et al. 2016).

318

319 Conclusion

Metagenomic exploration of wastewater in open-air canals, used for urban agriculture in Ouagadougou, revealed a wide functional specialization of microbial communities, a high prevalence of pathogen-specific and commonly found VFs, a variety of plasmid incompatibility types and multiple ARGs in this environment, suggesting that wastewater is a "hot spot" for formation and dissemination of antibiotic resistance. Our data showed that wastewater for urban agriculture in Burkina Faso might contribute to dissemination of bacterial resistance, including ESBLs and carbapenemases. That is of special concern for both community and hospital acquired infections. Further investigations are needed to determine the extent that exposed populations are

327	affected by this health issue. There is an urgent need to improve global access to clean water, sanitation, and
328	hygiene in LMICs to prevent the rise and dissemination of bacterial resistance from the environment to people.
329	
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