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Founder effects determine the genetic structure of the water flea Daphnia in Ethiopian reservoirs

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1	Founder effects determine the genetic structure of the water flea Daphnia in Ethiopian reservoirs
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17	Running headline: Founder effects in zooplankton populations

19 Abstract

20 Founder effects introduce stochasticity in the genetic structure of species at the regional scale. To 21 the extent that founder effects are important that they will result in a reduced signature of space, 22 time and environmental variation in landscape genetic data. We studied the metapopulation 23 genetic structure of recently founded populations of the microcrustacean Daphnia sinensis in ten 24 Ethiopian water reservoirs. We used three different approaches of estimating the number of 25 effective founders applied to two independent genetic marker sets to investigate the role of 26 founder effects and to estimate effective size of the founding population. Estimates of founding 27 sizes rarely exceeded eight individuals but were most often limited to less than four individuals. 28 No associations of genetic identities, gene frequencies, measures of genetic diversity or 29 differentiation with environmental and spatial variables were found. Age and size of the 30 reservoirs were not correlated with genetic diversity measures or number of founders in these 31 reservoirs. These findings indicate that neither strong selection, nor dispersal limitation are 32 responsible for the observed pattern of genetic variation. Our results suggest a regional 33 population structure that is strongly impacted by founder events, reflecting colonization by just a 34 few founders per water body, and not noticeably influenced by subsequent dispersal and gene 35 flow. Our results show that rapid colonization of empty habitats and fast population growth by a 36 handful of founders can result in strong founder effects, even in relatively large habitats 37 (estimated populations sizes of several million individuals) that are likely regularly reached by 38 new immigrants. 39 **Keywords**: colonization, *Daphnia sinensis*, effective population size, founder effects,

40 metapopulation, monopolization, zooplankton

42 Introduction

43 Metapopulation theory describes the interplay between colonization and extinction rates on patch
44 occupancy as a function of death, birth and dispersal rates (Hanski 1998; Levins 1969).

45 Population genetics, on the other hand considers occupied patches, and considers how gene flow

46 and population size interact to influence genetic structure within and among demes of a

47 metapopulation (Wright 1951). In reality, both extinction-colonization dynamics of local

48 populations and changes in genetic diversity by gene flow and drift within these local populations

49 act simultaneously. Dispersal, which is one of the most fundamental processes in ecology, affects

50 many aspects of evolution and population genetics if translated into successful colonization

51 (Bilton *et al.* 2001). Dispersal allows individuals to establish new populations in an empty patch

52 and promotes range expansion following colonization of new sites.

53 In a metapopulation genetics context, the colonization of an empty habitat patch by 54 founders can be considered as a special case of gene flow. As the new local population grows in 55 size to carrying capacity, new neutral immigrants (having equal expected fitness) can still enter 56 the population, but their relative contribution to the local gene pool is expected to decrease as 57 local population size at the time of immigration becomes larger. In this initial colonization 58 scenario at least two processes are responsible for successful colonization and establishment of a 59 population in an empty patch. First, the response of the immigrants to the local environmental 60 conditions of the habitat they colonize. Second, differences in their time of arrival at the site, 61 which generates a numerical advantage to the first colonizers over late-comers (Boileau *et al.* 62 1992; De Meester et al. 2002). In contrast to the relative ease of establishment of founders, which 63 experience no to a little competition, individuals attempting to immigrate into an established population close to carrying capacity are faced with strong intraspecific competition by the 64 65 resident population and low levels of resources. Due to this, realized rates of gene flow may be

much lower than expected based on the rates of dispersal (<u>De Meester *et al.* 2002</u>). This reduced
establishment success may strongly contribute to prolonged persistence of founder effects
(Boileau *et al.* 1992; <u>De Meester *et al.* 2002</u>; <u>Ventura *et al.* 2014</u>). Because founder effects
represent a type of sampling error, they introduce stochasticity in the genetic structure at the
regional scale, which tends to result in a reduced signature of space and environmental variation
in the genetic data (<u>Orsini *et al.* 2013</u>).

72 Estimates of the number of founders represent baseline estimates for ecological dispersal 73 rates (m_c the observed number of migrants), which represent the maximal potential for gene flow 74 (m_{e.} the effective number of migrants) among populations. Gene flow, the realized effect of 75 ecological dispersal on genetic structure, can be estimated indirectly through population genetics 76 as well (Broquet & Petit 2009). Although we know from many population genetic studies in 77 zooplankton that gene flow (m_e) is often much lower than expected, we have relatively few good 78 estimates of ecological dispersal rates (m_c) , because they are so hard to measure directly (Bilton 79 et al. 2001). Nevertheless, good dispersal estimates provide baseline information for a broad 80 array of ecological and evolutionary studies (Broquet & Petit 2009). Distinguishing between 81 dispersal and gene flow is essential, especially for biological conservation of populations and 82 species.

Here we take advantage of the recent creation of water reservoirs in Northern Ethiopia and the colonization of these water bodies by zooplankton to estimate the number of founders of populations of a zooplankton species using genetic methods. The reservoirs studied here are young (6-18 years) and two to three orders of magnitude larger than most other systems studied so far on founder effects in zooplankton and small invertebrates (Boileau *et al.* 1992; Haag *et al.* 2006; Louette *et al.* 2007). Specifically, we present patterns of genetic composition and differentiation of the water flea *Daphnia sinensis* in reservoirs that range in size from 1.8 to 45.4

90 hectare. Using variation at nuclear (nDNA) and mitochondrial (mtDNA) genetic markers, we 91 estimate allele frequencies, within-population genetic variation and among-population genetic 92 differentiation, and relate genetic variation and genotype composition to spatial, environmental 93 and temporal variables. We use various methods to independently estimate founding population 94 sizes, and thereby provide baseline estimates of dispersal rates. Using information on the 95 observed genetic structure (F_{ST}) and the associated expected gene flow at various levels of 96 migration-drift equilibrium and ages of populations, we show that the dispersal rates (m_c) are 97 orders of magnitude higher than the actual gene flow rates (m_e).

99 Methods

100 Study region and sampling

101 The studied reservoirs are part of a set of reservoirs constructed between 1984 and 2001 in the 102 highlands of Tigray Regional State, Northern Ethiopia. The rainfall in Tigray region is seasonal 103 and erratic resulting in moisture stress that hampers the rain-fed agriculture (Haregeweyn *et al.* 104 2006). To solve this problem agricultural development through irrigation has been a priority for 105 the Regional Government of Tigray. Hence, the target of the construction of reservoirs was 106 mainly to bring food self-sufficiency to the area through irrigation but also to use the water for 107 household consumptions (Asmelash et al. 2007). Thirty-two of these reservoirs have been the subject of a detailed limnological survey (Asmelash et al. 2007; Dejenie et al. 2008). Apart from 108 109 a single natural lake, not inhabited by the focal species of this study, *Daphnia sinensis*, no similar 110 large and deep aquatic systems are known from Tigray (Dejenie *et al.* 2008). Naturally, this 111 species occurs in temporary pools and ponds as well as larger temporary ponds and lakes (Gu et 112 al. 2013). The rapid colonization of these reservoirs shortly after their creation by a considerable 113 number of zooplankton taxa, including typical lake species (Dejenie et al. 2008), despite a 114 regional lack of similar habitats suggests that dispersal rates are relatively high and long-distance 115 dispersal events are rather frequent. Water birds (members of the Podicipidae, Pelecanidae, 116 Ciconiidae, Anatidae and Charadriidae family) are common in and alongside the reservoirs 117 (Asmelash et al. 2007) and are probably important vectors of dormant propagules of zooplankton 118 (Figuerola & Green 2002).

Thirty-two reservoirs were sampled for zooplankton in September 2005 (Dejenie et al.
2008). Ten of these samples contained *D. sinensis* in large enough numbers for population
genetic analyses (see Table S1, Supporting information). In addition, five temporary natural
wetlands were sampled, two of which contained *D. sinensis* (henceforth called T1 and T3),

123 bringing the total number of independent samples to twelve. All *Daphnia* samples were preserved 124 in 100% ethanol until further processing. Although previously identified as D. carinata King by 125 Dejenie et al. (2008), DNA barcoding indicates that individuals from these reservoirs belong to 126 D. sinensis, a member of the Daphnia similis species complex (Popova et al. 2016). 127 Measurements of geographic position and morphometric, physical, chemical, and biotic variables 128 were recorded for each sampled reservoir (see Table S1, Supporting information; Dejenie *et al.* 129 2008). Age of the reservoirs was expressed as number of years at sampling time since 130 construction of the reservoir. The two natural populations T1 and T3 were first excluded from all 131 age related analysis, and were in a second analysis arbitrarily given the same age as the oldest 132 reservoir. 133 134 Genotyping 135 DNA of individual *Daphnia* was extracted using the HotShot protocol (Montero-Pau et al. 2008) 136 Sample sizes ranged from 6 to 35 individuals per population for mtDNA (Table 1; 10 out of 12 137 samples with 17 or more individuals) and from 15 to 34 individuals per population for nDNA 138 (Table 2). Differences in sample sizes between both markers are due to unsuccessful 139 amplification with either approach. A fragment of 341 nucleotides of the mtDNA cytochrome 140 oxidase gene, subunit 1 (COI) was amplified using primers SCox1F1 (GGC CCC AGA TAT 141 GGC TTT) and SCox1R2 (GCT CCA GCT AAT ACT GGT AAA CTT), specifically designed 142 for this study. The polymerase chain reaction mix of 25 μ l contained 2 μ l DNA, 2.75 μ l 10x PCR 143 buffer (10 mM Tris-HCl; pH 8.3; 50 mM KCl), 0.4 µM of each primer, 2.2 mM MgCl₂, 0.2 mM 144 of each dNTP, and 1 unit Silverstar *Taq* DNA polymerase (Eurogentec[®], Liege Belgium). PCR 145 cycling conditions, PCR product purification and sequencing followed the methods of Mergeay et 146 al. (2007). The purified fragments were sequenced using 3.2 pmol of SCox1F1 primer and the

147 ABI Big Dye Terminator Kit. Sequences were aligned and trimmed in Mega 4.1 (<u>Kumar *et al.*</u>
148 2008).

149 Variation at six microsatellite loci using primers originally developed for the related 150 species Daphnia magna (B088, B172, B087, S6-38, B064 and Dma15 (Agostini et al. 2010; 151 Jansen B et al. 2011) was assessed in a single multiplex PCR reaction of 10 µl consisting of 5 µl HotStar *Taq* DNA polymerase buffer (Qiagen[®], Hilden Germany), 0.15 µM, 0.5 µM, 0.3 µM, 0.2 152 153 µM, 0.1 µM and 0.3 µM of each primer of locus Dma15, B087, B064, S6-38, B088 and B172, 154 respectively, and 2 µl of template DNA. Cycling conditions were 15' hot start denaturation at 155 95°C followed by 30 cycles of 30" for each step at 95°C, 56°C and 72°C, and a final elongation step at 60°C for 30'. Polymorphism was assessed on an ABI PRISM[®] 3130 genetic analyser 156 (Applied Biosystems[®], Foster City, CA, USA), using an internal Liz Gene-scan size standard by 157 means of the Genemapper 4.0 software (Applied Biosystems[®], Foster City, CA, USA). 158

159

160 **Population genetic data analysis**

161 Nucleotide diversity of the sequenced COI fragments was calculated in DNAsp version 162 4.5 (Rozas et al. 2003). Because we were not interested in the evolutionary relationships among 163 haplotypes that originated thousands of years prior to the colonization of these reservoirs, no 164 attempts were made to construct haplotype networks or to calculate genetic differentiation among 165 populations based on haplotype identity. The observed haplotype frequencies were primarily used 166 to estimate the number of founders involved in the colonization of each reservoir. We calculated 167 observed haplotype richness (HR) for each sample as well as haplotype diversity (HD). HD was 168 calculated as the true diversity equivalent of the Simpson concentration (Jost 2007). These values 169 were compared to expected HR and HD under the null hypothesis that all reservoirs form one panmictic population, using 10^4 permutations in Partition (Veech & Crist 2009). This yields alpha 170

171 (local gene diversity) and beta (average differentiation) estimators that are converted to their true172 diversity equivalents (Jost 2007).

173 Standard measures of genetic diversity (number of alleles, allelic richness per locus and 174 per population across all loci, observed heterozygosity and expected heterozygosity) at six 175 microsatellite loci were assessed in R using diveRsity package (Keenan et al. 2013). Identical 176 multilocus genotypes, on the basis of the combined information of six microsatellite loci, in a 177 given water body were considered to belong to a single clone. Clonal diversity (CD) was 178 expressed as the true diversity equivalent of the Simpson concentration, clonal richness (CR) as 179 the number of multilocus genotypes per water body. Moreover, relative clonal richness was 180 calculated per sample corrected for sample size expressed as proportion of clones to total 181 individuals genotyped as R = (G-1) / (N-1), where G is the number of genotypes and N indicates 182 sample size. We used HWclon (De Meester & Vanoverbeke 1999) to estimate whether or not 183 observed levels of CD and CR were significantly different from a random distribution, given the 184 genetic diversity and allelic polymorphism in the population (De Meester & Vanoverbeke 1999; 185 Vanoverbeke & De Meester 1997).

The standardized genetic variance among populations (F_{ST}) was calculated according to Weir & Cockerham(<u>1984</u>). We used 500 bootstrap pseudoreplicates to estimate 95% confidence intervals of the F_{ST} values. Genetic structure was assessed using the unbiased estimators of Nei & Chesser (<u>1983</u>) of the overall gene diversity (H_T) and subpopulation gene diversity (H_S).

190

191 Spatial, environmental and temporal correlates of genetic differentiation

192 To investigate the role of spatial and environmental variables separately and to

193 disentangle the unique contribution of each variable matrix to the genetic structure of the studied

194 populations we used redundancy analyses (RDA, a linear constrained ordination technique Dray

195 <u>*et al.* 2006</u>). In a multivariate variation partitioning analysis, the contributions of local 196 environmental predictors (n= 16 variables provided in Table S1, Supporting information) and 197 spatial predictors are tested by generating adjusted redundancy statistics (R^2_{adj}) in an RDA 198 analysis. A significant effect of environment would imply sorting of clones with different traits 199 and niches along environmental gradients.

200 Under a model of persistent founder effects, we expect genetic structure to be mainly 201 caused by chance events as dispersal is likely not limiting at the investigated spatial scale. Hence, 202 we expect to find at most a weak spatial genetic structure in the data (Orsini et al. 2013). To test 203 this, with the nDNA allele frequency data we performed a principle coordinates analysis (PCoA) 204 and then used the population loadings of the first six PCoA axes as dependent variables in a 205 distance-based redundancy analysis (db-RDA). In this RDA we attempted to explain the observed 206 genetic variation as a function of distance-based eigenvector maps (dbMEM) (Dray et al. 2006). 207 This analysis allows to find spatial patterns in the genetic data other than linear ones, making this 208 a more powerful approach with lower type II error rates than Mantel tests (Legendre & Fortin 209 2010). Five positive dbMEM eigenvectors were retained and were used as explanatory variables 210 in a forward selection procedure. Although this particular approach has the risk of identifying a 211 false positive spatial signal (see Blanchet *et al.* 2008), the double stop criterion of Blanchet *et al.* 212 (2008) is very conservative with regard to small datasets. Here we take a more liberal approach, 213 involving forward selection of spatial variables without prior testing of the overall spatial model, 214 to make sure that any lack of a detectable spatial signal is not due to the use of conservative 215 statistical methods.

In parallel, we performed a separate RDA relating the mtDNA data to the spatial data
(dbMEM). These mtDNA data were Hellinger-transformed to allow the use of linear regression
analyses in zero-inflated data (Legendre & Gallagher 2001).

219 Using non-parametric correlation (Spearman's rho) analyses, we related diversity 220 measures (He, AR, HR, HD, CR, CD, number of founders at mtDNA, and number of founders at 221 nDNA) to age, depth and size of the reservoirs. Under a model of persistent founder effects, 222 genetic diversity should not be related to age or size of the reservoir, as late-arriving immigrants 223 are expected to have little impact on the genetic structure compared to the very first founders. If 224 founder effects still persisted at the time of sampling, we expect that old and large populations do 225 not differ in genetic structure from younger and/or smaller populations. Specifically, we tested if 226 a model of population structure including reservoir age, reservoir size or their interaction 227 explained the genetic structure better than the null model not including age or size using Geste v. 228 2 (Foll & Gaggiotti 2006). Geste calculates a specific F_{ST} value for each population that 229 represents the population specific contribution to the total genetic differentiation in the 230 metapopulation. Then, the effect of age and/or size on these F_{ST} values is evaluated using generalized linear models. The posterior probability of each model was used to select the model 231 232 with the highest probability given our data. 233

234 Estimating the number of effective founders

We used three different approaches to estimate the number of effective founders, which reflects how many individuals contributed to the observed genetic diversity in each local population. First, we used a general approach based on F-statistics from microsatellite data, using the principle that the inbreeding coefficient among populations just after colonization is $F_{ST} = (2K)^{-1}$, with K the average number of founders per population (Boileau *et al.* 1992; Wade & McCauley 1988). Confidence intervals (95% CI) for F_{ST} were calculated by bootstrapping over loci (500 replicates). This method provides the average effective size of the founding population.

242 Second, we used a general simulation approach in the programming environment R (R 2.14; The 243 R Foundation for Statistical Computing, 2014) to calculate the expected HR and HD under a 244 model of random colonization from a regional gene pool with 1, 2... 10, 15 and 20 founders. Expected HR and HD values were calculated by randomly sampling 10⁴ times the pre-set number 245 246 of founders in 200 populations from an estimated regional frequency distribution, which was 247 based on the actual haplotype counts over all water bodies, or on presence-absence data for each haplotype per water body. For each of the random samples we calculated the probability that the 248 249 expected average HR or HD, over the 200 population, was smaller or larger than the values 250 observed in our empirical dataset. The product of these values provides the overall probability 251 that the observed HD or HR is achieved by the corresponding number of founders. The R-script 252 is available as supplementary information (Table S2, Supporting information), and provides the 253 average census size of the founding population. Third, we used a population-specific approach 254 for which we used the Colonize script (Mergeay et al. 2007; Vanoverbeke & Mergeay 2007). 255 This is a standalone command-line tool that calculates the likelihood that a predefined number of 256 founders from a predefined source population established the focal population, given the gene 257 frequency distribution of the source and the sink populations. The number of founders associated 258 to the highest likelihood score provides the best estimate for the founding propagule size of that 259 population. Again, this provides the census size of the founding population, which ignores that 260 some individuals contributed less to the genetic structure of the founding population than other 261 individuals. For each water body, we calculated likelihood scores for one to thirty founders, and 262 set both the number of batches and the number of random samples to 500. Ideally, one has 263 multiple putative source populations from which to sample, so as to assign the most likely source 264 population as well (see Mergeay et al. 2007, for an example). Here we have no such prior 265 information, and hence we use the regional gene pool (the average over all our samples) as the

266 overall source population. To increase the overall robustness of this approach, three different 267 prior allele frequency distributions were used. 1) Using the regional frequency of each allele over 268 the pooled data of all investigated water bodies (distribution = Freq.). Here we used the rare allele 269 correction in Colonize to account for extremely rare alleles. 2) Using presence-absence of each 270 allele per population and counting the frequency of occurrence of each allele over all populations 271 (distribution = Rich.). This approach gives less weight to alleles that dominate in certain 272 populations but are rare in other populations. 3) Using three abundance classes for the regional 273 allele frequencies (1: frequency <15%; 2: frequency 15-30%; 3: frequency >30%). This approach 274 (distribution = Level) gives even more weight to rare alleles. Analyses with Colonize were 275 performed separately for the mtDNA data and for the nDNA data in order to obtain independent 276 estimates for both marker types. Overall, these three approaches yield one overall estimate based 277 on F_{ST} (first method), two overall estimates based on mtDNA haplotype richness and haplotype 278 distributions (second method), and per reservoir three estimates based on mtDNA and three based 279 on nDNA (third method).

280

281 **Testing assumptions**

All of the outlined methods to estimate the number of founders rely on similar assumptions, but vary in their sensitivity to violations thereof. Here we outline how we tested for violations of the assumptions. First, we assume that genetic drift has not yet strongly affected allele frequency distributions (especially fixation or loss of alleles), given that they were founded at most 6 to 18 years before sampling. Second, we assume that all founders are genetically independent.

288

The main source of genetic drift in cyclical parthenogens like Daphnia is clonal selection

289 (reduction in clonal and genetic diversity as a result of selection among clones in the population) 290 (De Meester et al. 2006), which may erode genetic diversity considerably and thereby reduce our 291 estimates of founding population sizes. To assess whether clonal population structure (leading to 292 a similar signal as genetic bottlenecks) has affected our results, we performed a Spearman Rank 293 order correlation between clonal diversity (CD) and clonal richness (CR) versus the estimated 294 number of founders per population. This was performed for all estimates of the number of 295 founders, which have different sensitivities for common and rare alleles. Significant correlations 296 would reflect that clonal erosion affects our estimates of founding population sizes. Furthermore, 297 because drift reduces richness faster than diversity (Cornuet & Luikart 1996), founding 298 population size estimates based on richness should be lower than those based on diversity indices 299 if genetic drift is really important. Next, genetic drift affects mitochondrial genetic structure 300 stronger than nuclear genetic structure, because the effective population size at mitochondrial 301 genes is smaller (Hamilton 2011). Hence our founding number estimates should be lower when 302 using mitochondrial data if these were strongly influenced by genetic drift. We used the Student 303 t-test to test for differences among all these cases.

304 **Results**

305 *Genetic diversity*

306 Among a total of 285 sequences we found 25 polymorphic positions out of 299 nucleotide 307 positions in the CO1 gene fragment. This resulted in six distinct mitochondrial haplotypes. Two 308 of these haplotypes, H5 and H6, were singletons detected from Adi Gela and Adi Kenafiz, 309 respectively. Four haplotypes were common, with overall frequency of occurrence of 36.0, 33.0, 310 20.3 and 9.70 % for H2, H1, H3 and H4, respectively (Table 1 and Fig. 1). Most populations, 311 including both natural systems, were dominated by one or two haplotypes (average HD = 1.74). 312 Overall, the observed haplotype diversity or richness in a given population was always significantly lower (X^2 text, p < 0.0001) than the expected diversity or richness assuming a 313 314 panmictic regional metapopulation (Table 1). Pairwise nucleotide diversity among haplotypes 315 ranged from 0.003 to 0.047 (overall nucleotide diversity = 0.021). 316 For the microsatellite markers, we found an average of 5.3 ± 3.6 (\pm standard deviation) 317 alleles per locus over the whole metapopulation, whereas the mean allelic richness per locus was 318 2.58 ± 0.5 per population. The number of alleles per locus ranged from 2 to 11, with a total of 32 319 alleles scored over the six microsatellite loci combined. The total number of alleles observed 320 across all loci per population ranged from 11 to 20, with a mean allelic richness of 2.32 alleles 321 per locus (Table 2). The observed heterozygosity for the 12 relatively young Daphnia 322 populations ranged from 0.21 to 0.57 whereas the expected heterozygosity (H_e) ranged from 0.24 323 to 0.54 per population (Table 2 and Table S3, Supporting information). 324 In total, we found 183 unique multilocus genotypes (MLGs) out of 293 individuals 325 successfully genotyped. The majority of those MLGs (88%) were represented by a single 326 individual whereas 5% of the MLGs (n=10) were represented by two individuals. Only a small

number of MLGs (n = 18) was shared between reservoirs. The highest clonal richness (CR = 27) and clonal diversity (CD = 22.3) was observed for Gum Selasa (Table 2). The difference between the observed clonal richness/diversity and expected clonal richness/diversity was not statistically different at α = 0.05 for all the 12 populations studied, indicating that there is no substantial clonal erosion (Table 2).

All values of pairwise genetic differentiation (F_{ST}) were significant (p < 0.05). Nearby population pairs were not more related to each other than distant pairs (Table S4, Supporting information). The highest pairwise F_{ST} value was observed in the comparison between T1 and Adi Kenafiz ($F_{ST} = 0.585$), while the lowest pairwise F_{ST} value ($F_{ST} = 0.037$) was between Gereb Awso and Dibla (Table S4, Supporting information).

337 None of the RDA analyses yielded a model with one or more spatial or environmental 338 explanatory variables that could significantly (p < 0.05) explain the variation in the genetic data, 339 either in the distribution of the mtDNA haplotypes, or in the allele frequency data of the 340 microsatellite loci. Exclusion of the two natural systems did not affect the general pattern. Mantel 341 tests between pairwise genetic distance (Nei's genetic distance) and geographic distances or 342 environmental distances yielded correlation coefficients of r = -0.114 (p = 0.658) and r = 0.181 343 (p = 0.212), respectively, thus confirming the absence of any spatial trend in the genetic data 344 (Fig. 2).

346 Number of founders

Method 1: F_{ST} -based. The overall among-population fixation index (F_{ST}) was 0.237, with 95% confidence intervals (CI) ranging from 0.180 < F_{ST} < 0.342. Without T1 and T3, F_{ST} equalled 0.219 (95% CI: 0.169 < F_{ST} < 0.309). Since $F_{ST} \approx 1/2N$ at colonization, this reflects average effective founding population sizes of 2.3 individuals (95% CI: 1.6 < Ne < 3.0). Put differently, the average genetic diversity we observed corresponds to a mean effective founding population size of 1.6 to 3.0 individuals.

353

Method 2: Comparing observed to expected richness and diversity estimates. We found an average observed haplotype richness HR = 2.5 and an average observed haplotype diversity HD = 1.74 (Table 1). When comparing the average observed levels of HR and HD to expected HR and HD, we found that average founding population size estimates smaller than two and larger than eight are improbable at p-value = 0.05 (Table S5, Supporting information). The highest probability scores were obtained with 4, 3, 3 and 4 founders for the four types of simulations (Table S5, Supporting information).

361

362 Method 3: Population-specific simulations. We used population-specific simulations using 363 mtDNA and nDNA, based on three prior theoretical allele frequency distributions (Freq, Rich, 364 and Level, in descending order of sensitivity to rare alleles). For mtDNA, all three prior allele 365 frequency distributions yielded very comparable estimates (Table 3 and Table S6, Supporting 366 information), with averages ranging between 3.5 and 4 founders ($1 \le range \le 8$). This didn't 367 change appreciably when the natural systems were excluded (results not shown). For nDNA, 368 similar average (2.8 to 5.5) values were found ($2 \le \text{range} \le 13$), although the estimate using the 369 Freq. prior distribution was somewhat higher and was positively skewed due to higher estimates

for two populations (Table 3 and Table S6, Supporting information). Confidence in the prior
distribution of regional allele frequencies (Freq.) was unacceptably low (highest observed
likelihood score < 0.05) in six cases for nDNA and two cases for mtDNA. The prior distribution
based on local richness of alleles (Rich) yielded one estimate at nDNA with too low likelihood
scores (Table 3 and Table S6, Supporting information). All these estimates indicate that founding
population sizes were typically smaller than five individuals, and very rarely exceeded ten
individuals.

377

378 Effect of size and age of the reservoir

379 There was no significant relation (p > 0.05) between H_e and surface area or log (surface 380 area) of the water body (S = 183.28, Spearman rank r = 0.3563, p-value = 0.126) or with age of 381 the reservoir (S = 342.35, Spearman rank r = -0.197, p-value = 0.730, Figs S1, Supporting 382 information). The only significant correlation found in a total of 42 associations tested (age, 383 depth, area, versus AR, H_e, H_o, HR, HD, CR/N, CD/N, F_{IS}, number of founders at mtDNA 384 (Colonize-Rich), and number of founders at nDNA (Colonize-Rich)) was between H_e and 385 average depth (r = -0.79, p < 0.001). However, after Bonferroni correction, this p-value was 386 larger than 0.05. All other correlations were extremely weak and statistically insignificant at 387 α =0.05 (absolute value of r < 0.15, uncorrected p > 0.10; Figs S1, Supporting information). 388 Inclusion of size and/or age of the reservoir did not provide a better model (Geste v. 2) for 389 genetic structure than the more parsimonious null model. All of these results support the 390 hypothesis that founder effects are the main drivers of genetic structure and indicate that clonal 391 genetic drift did not markedly influence our estimates of founding population sizes. In addition, 392 none of the tests we did could show a significant difference between founding population size 393 estimates based on mitochondrial versus nuclear genetic data, or based on richness versus

- diversity estimates (all p values > 0.05). Thus, we have no indication that the assumptions
- 395 concerning genetic drift were violated (see Table S7 and Fig. S1, supporting information).

398 **Discussion**

399 Although we included a broad set of environmental and spatial variables, we found no 400 pattern with environmental variation, space and time (age) in the distribution of genetic variation 401 of the studied *Daphnia sinensis* populations inhabiting reservoirs in Tigray. Both the nuclear and 402 mitochondrial markers that we used are expected to behave neutrally. As such, we expected a 403 stronger signature of space than of environment. Still, a correlation with environmental variables 404 may result when particular haplotypes would hitchhike with particular genotypes or fixed allele 405 combinations that are favoured under certain conditions. Two studies on strictly asexual 406 zooplankton with comparable sample sizes and statistical power found clear environmental 407 and/or spatial structuring in their studies (Aguilera et al. 2007; Pantel et al. 2011) suggesting that 408 the lack of patterns in our dataset is not merely a consequence of insufficient statistical power. 409 Evolution-mediated priority effects, the key feature of the monopolisation hypothesis (De Meester et al. 2002; De Meester et al. 2016), are expected to be less important in asexual taxa 410 411 than in similar sexual taxa, due to a reduced ability for rapid local adaptation in the asexuals. As a 412 consequence, it is expected that fitness differences among dispersed clones will lead to a match 413 between environmental gradients and landscape genetic structure in asexual taxa, similar to 414 species sorting in communities (De Bie et al. 2012; Leibold et al. 2004). Conversely, if the 415 colonizing propagules of a sexual species harbours sufficient genetic variation to allow local 416 genetic adaptation, the increased fitness of resident populations may reduce establishment 417 success of new immigrants and thus reduce gene flow (De Meester et al. 2016). As a result, the 418 match between environmental and genetic variation is expected to be less strong in sexual than in 419 asexual species, which is the emergent pattern from our study on a cyclically parthenogenetic 420 Daphnia species and contrasts with the two aforementioned studies that focused on obligately 421 parthenogenetic Daphnia.

422 To investigate the number of founders typically involved in the colonization of new 423 moderately-sized freshwater systems (ranging from 1.8-45.4 ha in size), we used three different 424 approaches that rely on different test statistics with varying prior parameters, applied on two 425 independent sets of genetic markers. All approaches indicated that typically less than five 426 founders per habitat were responsible for the observed pattern of genetic diversity in the studied 427 reservoirs. Irrespective of whether we used estimates based on richness data or more detailed 428 frequency data, we obtained very similar estimates, showing that our results are robust to strong 429 allele frequency changes that may have occurred since colonization. Admittedly, the different 430 approaches we used all rely on similar assumptions, including an absence of genetic drift since 431 colonization, and genetic independence of each founder. The second assumption that the founders 432 are genetically independent from each other may have been violated to some extent. Birds, for 433 example, may disperse more than one dormant stage at the same time from a single source, 434 thereby introducing multiple related propagules. Especially results from mtDNA are expected to 435 be prone to such bias, given the much lower local and regional genetic variation compared to the 436 levels of variation found at nDNA. Estimated numbers of founders for nDNA and mtDNA were, 437 however, very similar.

438 We have detected high genetic differentiation among population (F_{ST} = 0.232) and no 439 isolation by dispersal limitation. This indicates low levels of gene flow among populations. 440 Furthermore, we failed to detect isolation-by-environment (IBE), which rules out the possibility 441 that sorting of genotypes along environmental gradients similar to species sorting in communities 442 (Leibold *et al.* 2004) might have driven the observed high genetic differentiation among 443 populations. Thus, our results support the idea that colonization dynamics in a newly created 444 metapopulation are strongly affected by founder effects exerted by a limited number of founding 445 genotypes. The founder effects observed here indicate that metapopulation and colonization

446 dynamics in this species resemble a lottery model (Sale 1977). In Sale's (1977) lottery model, 447 individuals compete for a limited number of discrete resources and once a resource is claimed, an 448 individual cannot be usurped from it. The classic lottery model was formulated at the community 449 level and with respect to microsites. However, it here acts at the level of genetic variants of a 450 species, and at the habitat level in a metapopulation. Populations are thus founded by a small 451 number of individuals from a varied array of regional sources. As long as a local genetic variant 452 persists (also if persistence is mediated through dormant stages; (Mergeay et al. 2007), the niche 453 space will continue to be occupied by these local variants, thereby pre-empting niche space for 454 immigrants. Several empirical studies focusing on colonization of novel habitats have shown that 455 dispersal rates in zooplankton are high (Cáceres & Soluk 2002; Jenkins & Buikema 1998; 456 Louette & De Meester 2005). The lack of spatial genetic patterns in our dataset also suggests that 457 dispersal per se is not limiting at the spatial scale here studied. 458 The sole environmental variable that showed a significant but negative correlation with 459 genetic diversity was average lake depth. One may speculate that the negative correlation 460 between depth and H_e reflects a species-specific preference for shallow waters, thereby reducing 461 the likelihood that a colonizing propagule will survive in deep reservoirs. This is indeed expected 462 from an organism that seems to naturally inhabit shallow pools. In that case, however, we would 463 also expect a similar negative relation between depth and number of founders, or other measures 464 of genetic diversity, which was not the case. An alternative explanation is that deeper lakes result 465 in more stable habitat conditions and therefore in populations that survive year-round and are 466 thus less dependent on dormant egg banks for survival. It is well known that more permanent

467 populations in *Daphnia* exhibit lower genetic diversity because of ongoing clonal erosion (<u>De</u>

468 Meester et al. 2006; Hebert 1987).

469

Earlier studies (Boileau et al. 1992; Haag et al. 2006) already showed that founder events

470 can strongly determine metapopulation structure, but the habitats they studied were very small 471 (less than $<100 \text{ m}^2$). The systems we study are thousand times larger than the typical size of the 472 small habitats studied earlier, with associated differences in carrying capacity, effective 473 population size, genetic drift and inbreeding. Although the results shown here should be 474 interpreted with some caution given that the limited number of reservoirs that was inhabited by 475 the studied *Daphnia* species resulted in a reduced statistical power in detecting spatial and 476 environmental patterns, our analyses strongly indicate that zooplankton populations of these new 477 large water bodies are typically founded by just a handful of individuals. Interestingly, the 478 number of founders in these reservoirs (on average 4-6) is strikingly similar to the range found in 479 ponds with population sizes that are up to a thousand times smaller (Boileau *et al.* 1992; Louette et al. 2007). Similarly, the local recolonization by *Daphnia barbata* of the 150 km² large Kenyan 480 481 Lake Naivasha happened most likely by no more than nine individuals from an old dormant egg 482 bank (Mergeay et al. 2007).

483 Inbreeding effective population size (N_e) in populations is a function of the number of 484 founders and is thus generally small in our zooplankton population. It seems that in zooplankton, 485 habitat size per se, at least within given boundaries, may have little influence on the effective 486 population size. Next to the low number of founders that seem typically involved, our results 487 indicate that these founder effects were equally high irrespective of the age of the reservoirs. 488 Several case studies on the propagule banks of *Daphnia* populations have demonstrated high 489 local genetic stability over periods of 50-150 years (Decaestecker et al. 2007; Mergeay et al. 490 2007). Recently, Ventura et al. (2014) even provided empirical evidence for founder effects 491 lasting thousands of years. All this evidence indicates that zooplankton populations primarily 492 have founder-controlled populations (Okamura & Freeland 2002), similar to founder-controlled 493 communities (Sale 1977). In such populations, dispersal contributes little to gene flow and is

494 mostly prevalent during the initial phase of colonization of empty or newly created habitats. 495 While dormant propagules are the main unit of dispersal in most zooplankton, their most 496 pervasive impact on landscape genetic structure may be their role in the short-term and long-term 497 local persistence of populations as well as in fostering colonization of empty habitats rather than 498 that they contribute to continuous gene flow among populations. Even seemingly extinct 499 populations may still be recolonized by local dormant egg banks once the habitat becomes 500 suitable again after decades (Mergeay et al. 2007). This has profound consequences for our view 501 on metapopulation biology of zooplankton and other micro-organisms, as these species often 502 share the lack of landscape genetic structure reflecting strong isolation-by-distance (Okamura & 503 Freeland 2002). More specifically, we should not equal high potential for dispersal into high rates 504 of gene flow (De Meester et al. 2016). In very small water bodies, however, negative effects of 505 genetic drift and inbreeding can be pronounced, and the positive influence on fitness of 506 immigrant alleles or genotypes from immigrants may then promote immigration and gene flow 507 (Ebert et al. 2002). One might therefore expect a shift from a gene flow dominated system in 508 extremely small populations (Ebert et al 2002) to metapopulations that are more strongly 509 dominated by local processes combined with extinction-recolonization dynamics in somewhat 510 larger systems such as the reservoirs studied here, shallow lakes and the sometimes much smaller 511 (approx. 100 m²) farmland ponds (De Meester *et al.* 2002; Louette *et al.* 2007; Vanoverbeke & 512 De Meester 1997).

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521 Data Accessibility

Sampling locations, raw environmental data for each reservoir and microsatellite genotype data is
 stored in in Dryad[®].

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642 List of tables

Table 1: Genetic diversity in mtDNA haplotypes: observed frequencies (expressed as fractions)

of each haplotype per water body and diversity descriptors. N: number of individuals extracted

645 per sampling site. HR: haplotype richness. HD-Si: true haplotype diversity measured with the

646 Simpson index. Average observed alpha diversity is the average observed within-sample

647 diversity weighted by sample size. Average expected alpha diversity gives the expected value of

648 HR or HD given a panmictic population over all water bodies, using 10,000 permutations. The

range of the expected values shows the lowest and highest value among all permutations over individuals. All observed values deviate significantly (p<0.0001) from expected values. True beta

651 diversity is calculated as gamma/alpha.

652

		Haplo	type n°						
Water body	Ν	1	2	3	4	5	6	HR	HD-Si
		0	0.71	0.06	0.18	0.06	0	4	1.86
Adi Kenafiz 18			0	0.39	0.56	0	0.06	3	2.16
Dibla	27	0.26	0.07	0.67	0	0	0	3	1.93
Gereb Awso	35	1.00	0	0	0	0	0	1	1.00
Gereb Mihiz	31	0.45	0.16	0.39	0	0	0	3	2.63
Gum Selasa	33	0.27	0.06	0.27	0.39	0	0	4	3.25
Haiba 6		0.83	0	0.17	0	0	0	2	1.38
Mai Leba 29		0.03	0.93	0	0.03	0	0	3	1.15
Meala 32		0.97	0.03	0	0	0	0	2	1.06
Tsinkanet	12	0	1.00	0	0	0	0	1	1.00
Temp 1 (T1) 22		0.14	0.36	0.50	0	0	0	3	2.49
Temp 3 (T3)	23	0	1.00	0	0	0	0	1	1.00
Overall frequence	0.33	0.36	0.203	0.097	0.005	0.005			
Average observe	ed (alpha)	0.38	0.30	0.19	0.12	0.01	0.01	2.5	1.74
Total diversity (gamma)							6	3.43
Average expecte	ed alpha							4	3.40
Range expected	alpha							3.7-4.2	3.01-3.67
True beta divers	ity							2.4	2.10

Table 2: Clonal and genetic diversity based on microsatellite loci (nDNA). N: sample size; n: number of individuals with complete

- 655 genotypic information (6 loci) on which calculations of clonal richness (CR) and clonal diversity (CD) were based. CR=clonal
- 656 richness; CD=clonal diversity.

Water body	Observe			rved			Expected ^{\$}							
	Ν	n	CR	CD	CR/n	CD/n	$CR \pm S.e$	$CD \pm S.e$	А	AR	Но	He	$\mathrm{HWE}^{\mathrm{F}}$	F _{IS}
AG	30	20	19	18.18	0.95	0.91	20.92±0.09	20.86±0.02	20	2.97	0.38	0.54	0.001	0.302
AK	30	20	15	11.11	0.75	0.56	12.04 ± 0.03	11.40 ± 0.05	11	1.76	0.39	0.34	0.335	-0.164
DIB	30	27	19	12.79	0.70	0.47	25.13±0.05	22.86 ± 0.08	15	2.29	0.41	0.32	0.085	-0.29
GA	32	31	13	8.50	0.42	0.27	21.35±0.06	16.10 ± 0.08	14	1.95	0.34	0.27	0.108	-0.242
GM	36	34	19	6.64	0.56	0.20	28.09 ± 0.10	25.05 ± 0.08	19	2.46	0.43	0.39	0.001	-0.106
GS	40	32	27	22.26	0.84	0.70	22.93±0.03	22.08 ± 0.05	18	2.63	0.38	0.47	0.000	0.19
HA	16	16	16	16.00	1.00	1.00	14.99 ± 0.01	14.98 ± 0.01	17	2.62	0.35	0.41	0.001	0.145
ML	29	23	22	21.16	0.96	0.92	23.68 ± 0.02	23.42±0.03	13	2.13	0.57	0.47	0.012	-0.19
MA	30	20	3	1.23	0.15	0.06	21.71±0.04	20.67 ± 0.06	13	1.85	0.51	0.29	0.000	-0.755
TS	18	15	11	7.76	0.73	0.52	14.99±0.01	14.98 ± 0.01	18	2.92	0.39	0.52	0.000	0.258
T1	26	23	15	8.97	0.65	0.39	18.37 ± 0.04	14.58 ± 0.06	14	2.10	0.21	0.24	0.272	0.138
T3	32	32	31	30.12	0.97	0.94	29.63±0.04	28.52 ± 0.06	14	2.13	0.45	0.47	0.335	0.032

657 CR/n and CD/n refers to clonal richness and diversity, respectively, corrected for sample size expressed as proportion of clones to total

658 individuals genotyped. A = number of alleles; Ar = allelic richness; H_0 = observed heterozygosity; H_e = expected heterozygosity; F_{IS} =

659 fixation index between individuals within local populations. [¥]The numbers are the p-value from a goodness of fit to HWE expectations

660 test using Fisher's exact test method. ^{\$}refers to the expected clonal richness (CR) and clonal diversity (CD) under Equilibrium using

randomisation tests implemented in Hwclon (<u>De Meester & Vanoverbeke 1999</u>) There is no significant difference (at $\alpha = 0.05$) between

662 Observed CR/CD and expected CR/CD values for all population comparisons

Table 3: Summary results of Colonize analyses with three prior allele frequency distributions

664 (Freq, Rich, Level; see main text for explanation), showing the most likely number of founders

665 for each population, based on either mtDNA or nDNA data, for each population and averaged

over all populations. Sd: standard deviation. Values with asterisk indicate that the likelihood

score was too low (p < 0.05) to represent a reliable estimate. Non-integer values represent the

average of shared highest scores.

669

	N° of founders with highest likelihood score (Colonize)								
	mtDN	A		nDNA					
Water body	Freq	Rich	Level	Freq	Rich	Level			
Adi Gela	4*	4	4	2*	2	2			
Adi Kenafiz	5*	8	7	7*	7	5			
Dibla	5	4	4	5	3	2			
Gereb Awso	1	1	1	5	2	2			
Gereb Mihiz	5	6	4	12.5	4	4			
Gum Selasa	8	8	7.5	13	5	4			
Haiba	2.5	2	2	3*	2	2			
Mai Leba	4	3.5	3	5	3	2			
Meala	2	2	2	2*	2	2			
Tsinkanet	1	1	1	3*	3*	4			
T1	5	5	5	3*	2	2			
T3	1	1	1	6	3	2			
Average	3.63	3.79	3.46	5.54	3.17	2.75			
Standard deviation	2.17	2.55	2.23	3.71	1.53	1.14			

670

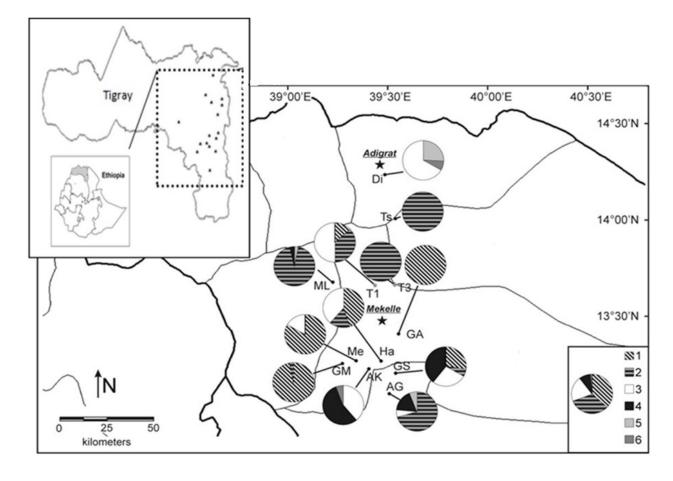






Figure 1: Geographic location of the sampling sites and mtDNA haplotype frequencies in each
population. Major cities are indicated with a star. Inset on the right shows the overall regional
frequency of the six encountered haplotypes. AG = Adi Gela; AK = Adi Kenafiz; Di = Dibla; GA
= Gereb Awso; GM = Gereb Mihiz; Ha = Haiba; ML = Mai Leba; Me = Meala; Ts = Tsinkanet;
T1 = Temporary pond 1; T3 = Temporary pond 3.

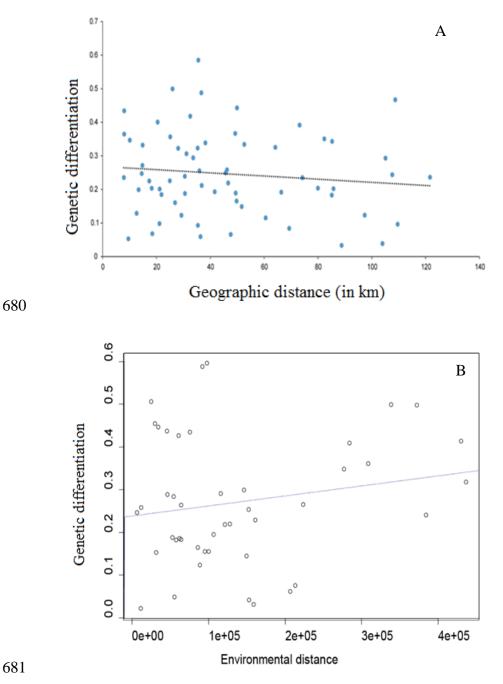


Figure 2. Relationship between Nei's genetic distance and geographic distance (panel A; testing for an isolation-by-distance and thus for dispersal limitation; r = -0.114; p = 0.662) and the Euclidean distance for environmental variables (panel B; testing for isolation-by-environment; r = 0.181; p= 0.212).